



Which ocean colour algorithm for MERIS in North West European waters?



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ABSTRACT

Chlorophyll-*a* (Chl *a*) is a key parameter for the assessment of water quality in coastal and shelf environments. The availability of satellite ocean colour offers the potential of monitoring these regions at unprecedented spatial and temporal scales, as long as a high level of accuracy can be achieved. To use satellite derived Chl *a* to monitor these environments, it is imperative that rigorous accuracy assessments are undertaken to select the most accurate ocean colour algorithm(s).

To this end, the accuracy of a range of ocean colour Chl *a* algorithms for use with Medium Imaging Resolution Spectrometer (MERIS) Level 2 (L2) Remote Sensing Reflectance (R_{rs}), using two different atmospheric correction (AC) processors (COASTCOLOUR and MERIS Ground Segment processor version 8.0 – MEGS8.0), were assessed in North West European waters. A total of 594 measurements of $R_{rs}(\lambda)$ and/or Chl *a* were made in the North Sea, Mediterranean Sea, along the Portuguese Coast, English Channel and Celtic Sea between June 2001 and March 2012, where Chl *a* varied from 0.2 to 35 mg m⁻³. The following algorithms were compared: MERIS Case 1 water Chl *a* algorithm OC4Me, the MERIS Case 2 algorithm Algal Pigment 2 (AP2), the MODIS-Aqua Case 1 Chl *a* algorithm OC3 adapted for MERIS (OC3Me), the MODIS-Aqua Garver-Siegel-Maritorena algorithm (GSM) adapted for MERIS and the Gohin et al. (2002) algorithm for MERIS (OC5Me). For both COASTCOLOUR and MEGS8.0 processors, OC5Me was the most accurate Chl *a* algorithm, which was within ~25% of *in situ* values in these coastal and shelf waters. The uncertainty in MEGS8.0 $R_{rs}(442)$ (~17%) was slightly higher compared to COASTCOLOUR (~12%) from 0.3 to 7 mg m⁻³ Chl *a*, but for $R_{rs}(560)$ the uncertainty was lower for MEGS8.0 (~10%) compared to COASTCOLOUR (~13%), which meant that MEGS8.0 Chl *a* was more accurate than COASTCOLOUR for all of the Chl *a* algorithms tested. Compared to OC5Me, OC4Me tended to over-estimate Chl *a*, which was caused by non-algal SPM especially at values > 14 g m⁻³. GSM also over-estimated Chl *a*, which was caused by variations in absorption coefficient of coloured dissolved organic matter at 442 nm ($a_{CDOM}(442)$). AP2 consistently under-estimated Chl *a*, especially when non-algal SPM was > 4 g m⁻³.

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1. Introduction

Information on marine environmental parameters, such as chlorophyll-*a* (Chl *a*), is fundamental for monitoring water quality, eutrophication and climate change (Birk et al., 2012; Boyce et al., 2010; Ferreira et al., 2011). Large scale spatial and temporal information on these parameters can be obtained by means of satellite remote sensing, which can aid our understanding of biogeochemical cycles (Chang et al., 2015; Siegel et al., 2014).

Monitoring the water quality of coastal and shelf waters is an integral part of water resource management, which allows tracking the effects of anthropogenic pollutants in the marine environment. Through the Urban Waste Water Treatment, Nitrate and Water Framework Directives, the European Union provided a general definition of 'good ecological status' for coastal waters based on 19 key parameters, which includes microbiological and physico-chemical variables. One of these parameters is Chl *a*, which is the photo-synthetically active pigment of phytoplankton, and can be measured indirectly from satellite ocean colour. In open ocean waters, a range of satellite ocean colour products have been developed, which have proven successful in areas where the principal optically active material in the water column is

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phytoplankton. It is however, more difficult to accurately determine Chl *a* from satellite in optically complex coastal and shelf regions (IOCCG, 2000). Coastal areas only account for ~7% of the world ocean's surface, yet play a globally important role in ameliorating human impacts on the marine ecosystem. Anthropogenic discharge of nutrients is advantageous for micro-phytoplankton such as diatoms and dinoflagellates, which potentially could result in increases in planktonic food web productivity in global coastal and shelf zones, and in turn may enhance annual carbon sequestration in marginal seas (Calbet et al., 2014). Coastal and shelf areas of Europe are commercially important for fishing and tourism, yet are subject to the increasingly adverse effects of harmful algal blooms (Baez et al., 2014; Glibert et al., 2014), eutrophication (Grizzetti et al., 2012; Romero et al., 2013) and climate change (Andersson et al., 2013; McQuatters-Gollop et al., 2007). In these regions, the presence of a_{CDOM} and mineral SPM as well as Chl *a* modify the light field (IOCCG, 2000), which makes accurate estimation of Chl *a* from satellite data in these regions, difficult. It is therefore necessary to develop and validate the accuracy of Chl *a* algorithms in coastal and shelf regions to facilitate monitoring of these environments.

The availability of data from the satellite sensor MERIS (2002–2012), which had more spectral bands and a higher spatial resolution compared to SeaWiFS (1997–2010) plus novel atmospheric correction (AC) models, has enabled the development of new products for optically complex waters. These include an algorithm that retrieves inherent optical properties (IOP) and biogeochemical parameters from Case 2 waters (Doerffer and Schiller, 2007). Directional water leaving radiance is input to the algorithm and it outputs Chl *a*, SPM and the absorption coefficient of detrital material and gelbstoff (a_{dg}) based on the conversion of scattering and absorption coefficients using non-linear multiple inversion solutions and regional conversion factors to give concentrations. This algorithm has been shown to be accurate in some (Doerffer and Schiller, 2007; Schiller and Doerffer, 2005), but not all Case 2 coastal (Cui et al., 2014; Melin et al., 2007; Tilstone et al., 2012), estuarine (Ambarwulan et al., 2010), freshwater lakes (Binding et al., 2011; Palmer et al., 2015) and other highly absorbing water bodies (Beltran-Abaunza et al., 2014; Folkestad et al., 2007; Ohde et al., 2007). As a consequence, a number of alternative algorithms have been proposed (Beltran-Abaunza et al., 2014; Hokedal et al., 2005; Peters et al., 2005; Tilstone et al., 2012; van der Woerd and Pasterkamp, 2004; Zibordi et al., 2009a). To improve satellite ocean colour algorithms, it is fundamental to improve the accuracy satellite R_{rs} that are used to calculate Chl *a*. To this end, ESA funded a number of initiatives to improve the AC scheme for MERIS which included improvements in the neural network (NN) AC model for coastal waters (COASTCOLOUR), and re-processing of the MERIS archive using the MEGS8.0 processor which uses both a bright pixel correction for highly scattering waters and a clear water AC model (Lerebourg et al., 2011). It is important to test these and to select the most accurate algorithm for remotely sensed Chl *a* in European coastal and shelf seas to coincide with the launch of the Copernicus new generation satellite Sentinel-3 (Donlon et al., 2012).

In this paper we validate a range of MERIS Chl *a* products, using both MEGS8.0 and COASTCOLOUR processors, for the North West European regions of the North Sea, Celtic Sea, English Channel, South and Western Iberian Peninsula and Mediterranean Sea. Five ocean colour algorithms were compared and the choice of the algorithms was based on their availability to the user community as standard products for MERIS and MODIS-Aqua.

2. Methods

2.1. Study area characteristics and sampling regime

Remote Sensing Reflectance ($R_{rs}(\lambda)$) and/or Chl *a* were measured at 594 stations between April 2001 and March 2012 in the North Sea, English Channel and Celtic Sea (in the figures English Channel also

refers to the Celtic Sea), Mediterranean Sea and along the Iberian coast (Fig. 1A). These data are available from ESA's MERMAID and COASTCOLOUR data sets (Nechad et al., 2015).

The North Sea is characterized by the entrance of warm, high saline water through the Orkney-Shetland inflow (off NE Scotland) and through the English Channel-Straits of Dover in the south (Fig. 1A), which drives a cyclonic pattern of circulation, which can cause re-suspension of sediment over the shallow North Sea shelf (Reid et al., 1988). The Portuguese coast is affected by seasonal upwelling from March to September and the poleward current during the winter months (Fiuza et al., 1982). In addition, the coastal waters close to Lisbon are strongly affected by the River Tagus, which elevates the CDOM and SPM on the shelf break (Valente and Da Silva, 2009). The Mediterranean Sea stations are in the central Ligurian Sea which is characterized by a cyclonic circulation that causes strong flows close to the coast which circulate north of the Island of Corsica and return in a south west direction (Fig. 1A) to establish a front between the coastal shelf and the deeper offshore water (Antoine et al., 2008).

North Sea and English Channel coastal areas have high absorption and scattering properties (Hommersom et al., 2009), which are optically variable due to different regional and seasonal contributions of a_{CDOM} , absorption coefficient of non-algal particles (a_{NAP}) and particulate back-scattering coefficient (Babin et al., 2003a; Babin et al., 2003b) and can therefore switch between case 1 and 2 water types (Groom et al., 2009). There are three main groups of specific-absorption properties in the North Sea, English Channel and Celtic Sea with low Chl *a* specific absorption coefficient of phytoplankton (a_{ph}^*) ($<0.03 \text{ mg m}^{-2}$), high SPM specific absorption coefficient of non-algal particles (a_{NAP}^*) ($>0.04 \text{ g m}^{-2}$) and a_{CDOM} ($>0.4 \text{ m}^{-1}$) close to the coast; medium a_{ph}^* ($>0.03, <0.05 \text{ mg m}^{-2}$), a_{NAP}^* ($>0.02, <0.04 \text{ g m}^{-2}$) and a_{CDOM} ($>0.1, <0.4 \text{ m}^{-1}$) over the shelf; and high a_{ph}^* ($>0.05 \text{ mg m}^{-2}$), low a_{NAP}^* ($<0.02 \text{ g m}^{-2}$) and a_{CDOM} ($<0.1 \text{ m}^{-1}$) further offshore (Tilstone et al., 2012). Portuguese coastal waters are dominated by a_{ph} , though a_{CDOM} can account for between 33 and 60% of the total absorption between bloom and non-bloom conditions (Goela et al., 2015). The principal satellite validation station in the Mediterranean Sea is BOUSSOLE, which is a deep clear water site where the dominant optically substance is phytoplankton. Other stations are sampled closer to the coast on the monthly transect to BOUSSOLE, which can be affected by other optically active substances (Antoine et al., 2008).

2.2. Measurement of normalized water leaving radiance and Remote Sensing Reflectance

Measurements of normalized water leaving radiance (nL_w) were performed by the Royal Belgian Institute of Natural Sciences (RBINS), Centre for Marine and Environmental Research (CIMA) and Laboratoire d'Océanographie de Villefranche (LOV). RBINS used three TriOS-RAMES hyperspectral spectro-radiometers, two measuring radiance and one measuring downwelling irradiance, mounted on a steel frame fixed to the bow of the ship, facing forward to minimize ship shadow and reflection, so that zenith angles of the sea and sky viewing radiance sensors were 40° as in (Ruddick et al. (2006)). To reduce sun glint and bi-directional reflectance effects, the ship was maneuvered on station to point the radiance sensors at a relative azimuth angle of 135° away from the sun. The spectro-radiometers were calibrated before and after the cruise consistent with SeaWiFS protocols (Mueller, 2000). Water-leaving reflectance (ρ_w) was calculated from simultaneous above-water measurements of downwelling irradiance (E_d^{0+}), total upwelling radiance (L_{sea}^{0+}) and sky radiance (L_{sky}^{0+}) (in the direction of the region of sky that reflects into the sea viewing sensor), from:

$$\rho_w(\lambda) = \pi \frac{L_{sea}^{0+}(\lambda) - \rho_{sky} L_{sky}^{0+}(\lambda)}{E_d^{0+}(\lambda)} \quad (1)$$

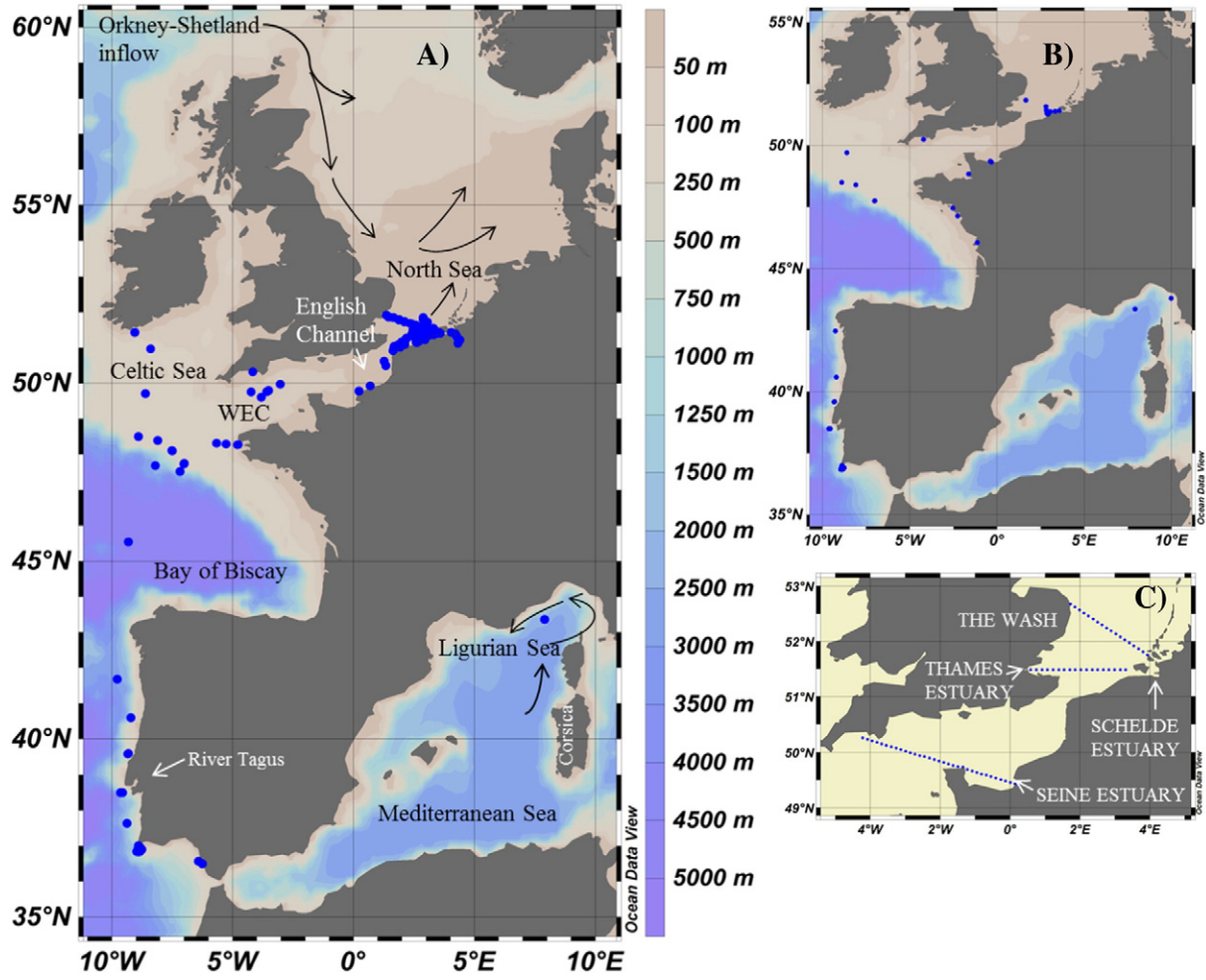


Fig. 1. Location of sampling stations for (A.) *in situ* measurements of R_{rs} and Chl *a* (B.) match-ups between *in situ* and MERIS MEGS8.0 & COASTCOLOUR R_{rs} and Chl *a*, (C.) Transects from which MEGS8.0 Chl *a* processed using OC5Me, OC4Me, AP2 and non-algal SPM were extracted. The colour scale indicates the bathymetry depth. In (A.), the arrows represent the predominant currents and illustrate the inflow of Atlantic water through the Orkney-Shetland channel in the North and the English Channel in the South and the flow of the Mediterranean Sea current around Corsica Island.

where ρ_{sky} is the air-water interface reflection coefficient for radiance. In the case of a flat sea surface ρ_{sky} is equal to the Fresnel reflection coefficient and is assumed to be 0.02 for clear skies for satellite ‘match-ups’ (Ruddick et al., 2006). Residual skylight was removed using baseline correction following Mobley (1999). The normalized water leaving radiance (nL_w) was calculated from:

$$nL_w(\lambda) = \frac{\rho_w(\lambda)}{\pi} \times f_0(\lambda) \text{ (in } \text{mW cm}^{-2} \mu\text{m}^{-1} \text{sr}^{-1}\text{)} \quad (2)$$

where $f_0(\lambda)$ is the mean solar flux above the earth's atmosphere. $R_{rs}(\lambda)$ (sr^{-1}) was then calculated from:

$$R_{rs}(\lambda) = \frac{L_w(\lambda)}{E_s(\lambda, 0^+)} \quad (3)$$

where $E_s(\lambda, 0^+)$ is the above surface downwelling spectral irradiance ($\text{W m}^{-2} \text{nm}^{-1}$) and $L_w(\lambda)$ is the water leaving radiance ($\text{W m}^{-2} \text{nm}^{-1} \text{sr}^{-1}$). NASA ocean optics protocols were used to compute $L_w(\lambda)$, $nL_w(\lambda)$ and $R_{rs}(\lambda)$ (Mueller, 2000). Surface downwelling irradiance ($E_s(\lambda, 0^+)$) was calculated from:

$$E_s(\lambda, 0^+) = (1 + \alpha)E_d(\lambda, 0^-) \quad (4)$$

where α is the Fresnel reflection albedo from air + sky (~ 0.043), and $E_d(\lambda, 0^-)$ is extrapolated from the $E_d(\lambda, z)$ profile.

CIMA collected *in situ* radiometric measurements at three sampling stations off South West Portugal, following MERIS protocols (Barker, 2011; Doerffer, 2002). Measurements of ρ_w were acquired with a Satlantic tethered attenuation coefficient chain sensor (TACCS), which consists of a floating buoy encasing a hyperspectral surface irradiance sensor ($E_s(\lambda)$) and a subsurface radiance sensor ($L_u(\lambda)$) located 0.62 m below the surface and a tethered attenuation chain supporting four subsurface irradiance sensors $E_d(z)$ at 2, 4, 8, and 16 m. Further details of the processing of E_s , L_u and E_d are given in (Cristina et al., 2014; Cristina et al., 2009). The ρ_w acquired by the TACCS was calculated from:

$$\rho_w(\lambda) = \pi \frac{L_w(\lambda)}{E_s(\lambda)} \quad (5)$$

where ρ_w is equivalent to R_{rs} (Eq. (3)) when scaling by a factor of π (i.e. $R_{rs} = \rho_w/\pi$).

Measurements taken by LOV were obtained from the radiometric data buoy BOUSSOLE using the methods described in Antoine et al. (2008).

2.3. Measurement of Chlorophyll-*a*

On all cruises, surface water samples were collected using 10 L Niskin bottles and between 0.25.

between 0.25 and 2 L of seawater were filtered onto 0.7 μm GF/F filters. PML, MARE and CIMA used High Performance Liquid Chromatography (HPLC) to determine Chl *a*. PML extracted phytoplankton pigments into 2 mL 100% acetone containing an internal standard (apocarotenolate; Sigma) using an ultrasonic probe (35 s, 50 W). Extracts were centrifuged to remove filter and cell debris (3 min at 20 xg) and analysed by HPLC using a reversed phase C8 column and gradient elution (Barlow et al., 1997) on an Agilent 1100 Series system with chilled autosampler (4 °C) and photodiode array detection (Agilent Technologies). CIMA extracted the pigments in 90% acetone for 4–6 h in a refrigerator (−4 °C) and using an ultrasonic probe for 20 s before and after refrigeration. Extracts were centrifuged and then analysed either on a Waters 600E HPLC system with diode array detector and a C8 Thermo Hypersil-Keystone (ODS-2) column or on an Agilent HPLC with diode array using a C8 Alltech Altima column following Goela et al. (2014). MARE extracted pigments with 2–5 mL of 95% cold-buffered methanol (2% ammonium acetate) for 30 min to 1 h at −20 °C, in the dark. Pigment extracts were analysed using a Shimadzu HPLC comprising a solvent delivery module (LC-10ADVP) with system controller (SCL-10AVP), a photodiode array (SPD-M10ADVP), and a fluorescence detector (RF-10AXL). Chromatographic separation was carried out with a C8 column, following Zapata et al. (2000). PML, CIMA and MARE calibrated the HPLC system using a suite of standards (DHI Water and Environment, Denmark and pigments in samples identified from retention time and spectral match using photodiode array spectroscopy (Jeffrey et al., 1997). Chl *a* concentration was calculated using response factors generated from calibration using a Chl *a* standard (DHI Water and Environment, Denmark). Samples taken by IFREMER along the North France coast were measured by spectrophotometry using the methods outlined in Lorenzen (1967).

2.4. Measurement of suspended particulate material (SPM)

All laboratories filtered between 0.5 and 3 L of seawater onto 47 mm, 0.7 μm GF/F filters in triplicate. The filters were ashed at 450 °C, washed for 5 mins in 0.5 L of MilliQ, and then dried in a hot air oven at 75 °C for one hour, pre-weighed and stored in desiccators (Van der Linde, 1998). Seawater samples were filtered in triplicate onto the pre-prepared filters which were then washed (including the rim) three times with 0.05 L MilliQ to remove residual salt. Blank filters were also washed with MilliQ to quantify any potential error due to incomplete drying. The filters were then dried at 75 °C for 24 h and weighed on microbalances (detection limit 10 μg). SPM concentrations were determined from the difference between blank and sample filters and the volume of seawater filtered.

2.5. Measurement of CDOM absorption coefficients ($a_{CDOM}(\lambda)$)

All laboratories filtered replicate seawater samples through 0.2 μm Whatman Nuclepore membrane filters into acid cleaned glassware. The first two 0.25 L of the filtered seawater were discarded. $a_{CDOM}(\lambda)$ was determined using the third sample in a 10 cm quartz cuvette

from 350 to 750 nm relative to a bi-distilled MilliQ reference blank and was calculated from the optical density of the sample and the cuvette path length following the protocols given in Tilstone et al. (2004).

2.6. Satellite data and algorithms

MERIS full resolution (300 m) COASTCOLOUR v2013 and reduced resolution (1 km) MEGS8.0 data were downloaded from Brockman Consult and extracted using Beam v4.8. Both MEGS8.0 and COASTCOLOUR L2 R_{rs} were used to process five Chl *a* algorithms. MEGS8.0 processor has a two-step approach to atmospheric correction; if there is any signal in the NIR due to backscattering by sediment or coccolithophores, the Bright Pixel AC (Moore and Lavender, 2010) is triggered (Lerebourg et al., 2011). If this is not triggered, the Clear Water Atmospheric Correction model is implemented (Antoine and Morel, 1997). The COASTCOLOUR AC is an NN based inversion technique that implements a forward model configured with a set of regional, coastal aerosol optical properties (Brockmann, 2011) collected from the global AERONET network (Zibordi et al., 2009b).

The following MERIS quality flags were used to eliminate erroneous data generated from the processors: cloud flag over ocean (CLOUD), land (LAND), no glint correction applied – accuracy uncertain (HIGH_GLINT), reflectance corrected for medium glint – accuracy maybe degraded (MEDIUM_GLINT), highly absorbing aerosols (AODB), low sun angle (LOW_SUN), low confidence flag for water leaving or surface reflectance (PCD1_13) and reflectance out of range (PCD_15). The MERIS L2 products were extracted from a 3×3 pixel box, within ± 0.5 h of MERIS overpasses. The functional form of each of the Chl *a* algorithms tested is given in Table 1 and described in brief below.

The first Chl *a* algorithm tested was the standard Case 2 MERIS Chl *a* product (AP2) (Doerffer and Schiller, 2007; Schiller and Doerffer, 2005), which uses an NN to derive the a_{ph} and the particulate scattering coefficient (b_p), and through empirical bio-optical relationships, Chl *a*, SPM and a_{dg} are computed. An inverse model solves the IOPs from $R_{rs}(\lambda)$ using a Look Up Table. The total absorption coefficient (a_{tot}) is then apportioned into a_{ph} and a_{dg} and empirical solutions are used to convert a_{ph} to Chl *a* and the b_{bp} to SPM. The AP2 algorithm was calibrated using a global dataset, which included a large IOP data set from North Sea coastal waters of the German Bight. AP2 solves Chl *a* from a_{ph} as follows; Chl *a* = $21.0 * a_{ph}(442)^{1.04}$ (Table 1).

The second algorithm tested is the default MERIS Case 1 Chl *a* algorithm (AP1; also known as OC4Me). OC4Me is the MERIS algal pigment 1 fourth-order polynomial algorithm, that was parameterised using a theoretical model (Morel and Maritorena, 2001), and tuned using *Kd* and Chl *a* data given in Morel and Antoine (2011). It uses the maximum Remote Sensing Reflectance ratio from: R_{rs} 442/560, R_{rs} 490/560 or 510/560. The switch between R_{rs} 442/560 and R_{rs} 490/560 theoretically occurs at a Chl *a* concentration of 0.534 mg m^{-3} and between R_{rs} 490/560 and R_{rs} 510/560 at 2.23 mg m^{-3} . Chl *a* was estimated using the equation given in Table 1, where ρ is the maximum $R_{rs}(\lambda)/560$ ratio.

Table 1
Functional form of MERIS Chl *a* algorithms.

Algorithm	Reference	Functional form
AP2	Doerffer and Schiller (2007)	$Chl\ a = 21.0 * a_{ph}(442)^{1.04}$
OC4Me	Morel and Antoine (2011)	$Chl\ a = 10^{(a + bR + cR^2 + dR5 + eR4)}$ where $R = \max\ \text{ratio of } \log_{10} \left\{ \left(\frac{R_{rs,443}}{R_{rs,560}} \right), \left(\frac{R_{rs,490}}{R_{rs,560}} \right), \left(\frac{R_{rs,510}}{R_{rs,560}} \right) \right\}$ $a = 0.4502748$; $b = -3.259491$; $c = 3.52271$; $d = -3.359422$; $e = 0.949586$.
OC3Me	O'Reilly et al. (2000)	$Chl\ a = 10^{(a + bR + cR^2 + dR5 + eR4)}$ where $R = \max\ \text{ratio of } \log_{10} \left\{ \left(\frac{R_{rs,442}}{R_{rs,560}} \right), \left(\frac{R_{rs,490}}{R_{rs,560}} \right) \right\}$ $a = 0.40657$; $b = -3.6303$; $c = 5.44357$; $d = 0.0015$; $e = -1.228$.
GSM	Maritorena et al. (2002)	$L_{wn}(\lambda) = \frac{t_{F1}(\lambda)}{t_w} \sum_{i=1}^2 g_i \left\{ \frac{[b_{bw}(\lambda) + b_{bp}(442)(\lambda/442)^{-1.0337}]}{[b_{bw}(\lambda) + b_{bp}(442)(\lambda/442)^{-1.0337}] + [a_w(\lambda) + Chl\ a * \sigma_{ph}(\lambda) + a_{CDOM}(442) * e^{-0.0206(\lambda - 442)}]} \right\}^i$
OC5Me	Gohin et al. (2002)	$Chl\ a = F(\text{triplet})$; where triplet = {max of the reflectance ratios as in OC4Me, nLw(412), nLw(560)}

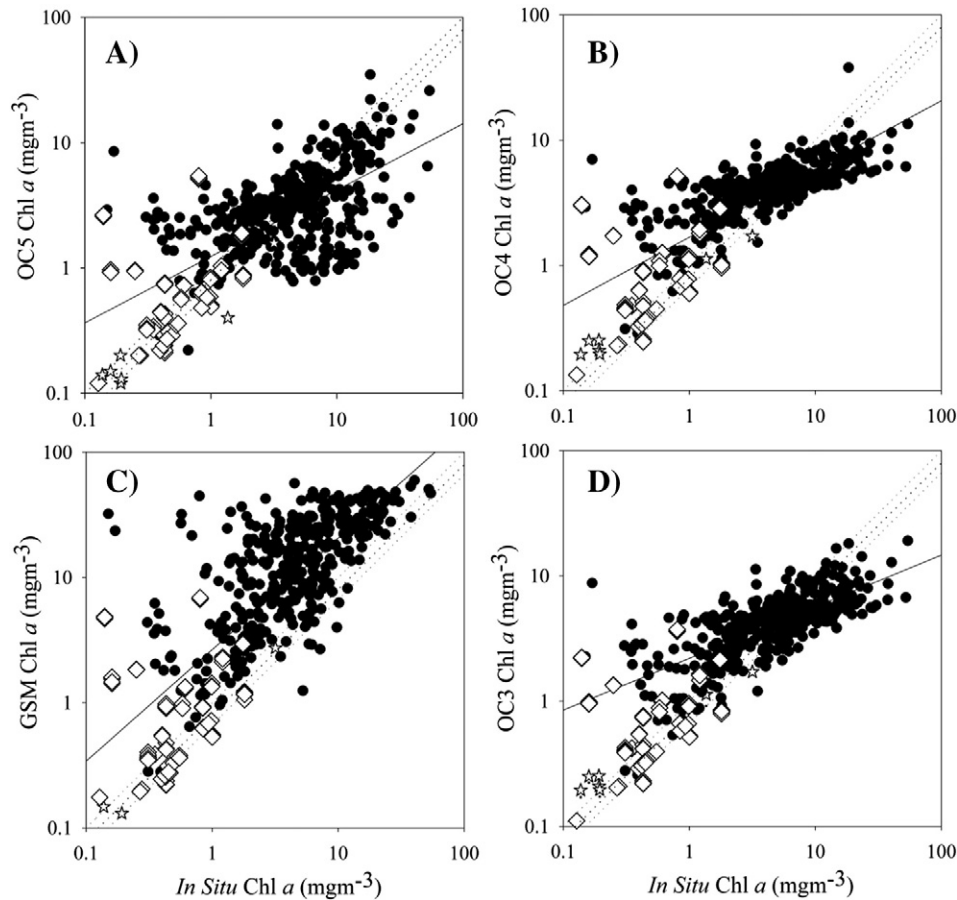


Fig. 2. Comparison of *in situ* Chl *a* and *in situ* R_{rs} derived Chl *a* for (A) OC5, (B) OC4, (C) GSM, (D) OC3. Faint dotted lines are the 1:1 line, upper and lower 20% quartiles. Solid line is the regression line. Filled circles are North Sea data, open diamonds are Portuguese coast, open stars are Mediterranean coast.

OC3Me is an empirical fourth-order band ratio algorithm that uses one of two $R_{rs}(\lambda)/R_{rs}(560)$ ratios; either $R_{rs}(442)/R_{rs}(560)$ or $R_{rs}(490)/R_{rs}(560)$, depending on the reflectance characteristics of the water type (O'Reilly et al., 2000). It is the default Case 1 water algorithm for MODIS-Aqua and has been adapted for MERIS using the maximum

Remote Sensing Reflectance ratio from: $R_{rs} 442/560$ or $R_{rs} 490/560$ and the coefficients for OC3Me adapted for MERIS given in Table 1.

OC5Me is a modified version of OC4Me, which includes an empirical parameterisation of 412 and 560 nm channels for MERIS related to the absorption of CDOM and scattering of SPM (Novoa et al., 2012;

Table 2

Performance indices for relative errors between *in situ* and modelled Chl *a* from AP2, GSM, OC3Me, OC4Me and OC5Me using *in situ*, COASTCOLOUR and MEGS8.0 $R_{rs}(\lambda)$. Percentage variance explained (R^2), intercept and slope and log-difference errors in measured and satellite Chl *a* ratio (r) as Mean (M), Standard deviation (S), root-mean square (Log_{10} -RMS) and root-mean square error (RMS-E). The geometric mean and one-sigma range of the ratio ($F = \text{Value}_{\text{alg}} / \text{Value}_{\text{meas}}$) are given by F_{med} , F_{min} , and F_{max} , respectively; values closer to 1 are more accurate. RPD is the relative percentage difference. The algorithm with the highest Chl *a* accuracy is highlighted in bold.

$R_{rs}(\lambda)$	R^2	Slope	Intercept	r	RPD	Log_{10} -RMS	M	S	F_{med}	F_{max}	F_{min}	RMS-E
In situ N = 482. (North Sea, N = 384; Portugal, N = 87; Med, N = 12).												
OC3Me	0.41	0.23	3.07	0.92	-9	0.23	-0.03	0.23	0.93	1.57	0.55	4.97
OC4Me	0.50	0.24	2.54	0.76	24	0.24	-0.02	0.24	0.95	1.66	0.55	5.77
GSM	0.47	1.34	6.41	1.13	-61	0.43	0.29	0.31	1.96	4.01	0.96	27.20
OC5Me	0.38	0.32	1.54	0.80	104	0.33	-0.11	0.28	0.78	1.49	0.41	5.90
COASTCOLOUR N = 113 (North Sea, N = 18; WEC, N = 16; Portugal, N = 66; Med, N = 13).												
AP2	0.61	0.71	0.61	1.42	83	0.44	0.03	0.43	1.07	2.92	0.39	2.46
OC3Me	0.65	0.72	0.26	2.05	64	0.42	0.05	0.42	1.12	2.92	0.43	2.10
OC4Me	0.67	0.76	0.46	0.68	20	0.43	-0.09	0.42	0.81	2.14	0.31	2.04
GSM ^a	0.61	0.99	0.10	2.42	105	0.40	0.05	0.54	1.13	3.89	0.33	2.01
OC5Me	0.75	0.65	0.26	1.58	63	0.39	0.06	0.39	1.14	2.79	0.44	2.14
MEGS8.0 N = 88 (North Sea, N = 18; WEC, N = 9; Portugal, N = 45; Med, N = 16).												
AP2	0.31	0.42	0.72	1.42	103	0.47	0.09	0.46	1.24	3.56	0.43	2.41
OC3Me	0.51	0.64	0.48	1.32	42	0.32	0.04	0.32	1.11	2.29	0.54	1.98
OC4Me	0.55	0.83	0.63	0.62	25	0.32	-0.09	0.31	0.81	1.65	0.40	2.14
GSM ^b	0.58	0.60	0.40	1.51	84	0.36	0.11	0.34	1.30	2.85	0.59	2.54
OC5Me	0.87	0.62	0.18	1.05	46	0.27	0.10	0.25	1.25	2.21	0.71	1.61

^a For GSM Chl *a* calculated using COASTCOLOUR R_{rs} , N = 89 match-ups were available (North Sea, N = 11; WEC, N = 0; Portugal, N = 65; Med, N = 13).

^b For GSM Chl *a* calculated using MEGS8.0 R_{rs} , N = 81 match-ups were available (North Sea, N = 18; WEC, N = 2; Portugal, N = 45; Med, N = 16).

Saulquin et al., 2011). Chl *a* concentration is determined from the triplet values of OC4Me maximum band ratio, $nL_w(412)$ and $nL_w(560)$, from a Look Up Table (LUT), based on the relationships between measured Chl *a* and satellite $R_{rs}(\lambda)$ from observations in the English Channel and Bay of Biscay (Gohin et al., 2002). The method has also been extended to the Mediterranean Sea and applied to MODIS-Aqua and MERIS (Gohin, 2011). OC5 satellite images using both MODIS-Aqua and MERIS are available from <https://www.neodaas.ac.uk/multiview/pa/>

The GSM is an optimized semi-analytical algorithm that simultaneously retrieves Chl *a*, a_{dg} and b_{bp} at 442 nm, from spectral measurements of $nL_w(\lambda)$. The parameters for the model were obtained

through simulated annealing which is a global optimization technique (Maritorena et al., 2002). The model was originally parameterised for SeaWiFS and in this paper has been adapted for MERIS using the amoeba implementation from SeaDAS with an update of the absorption spectrum values for pure water (Pope and Fry, 1997) to (Kou et al., 1993) which extend the spectrum beyond 700 nm to 709, 753 & 778 nm for MERIS.

The concentration of non-algal SPM (the inorganic component of SPM that is not related to phytoplankton) was derived using the method developed in Gohin et al. (2005). Once Chl *a* has been determined using OC5Me, non-algal SPM is estimated from $nL_w(560)$ or $nL_w(670)$

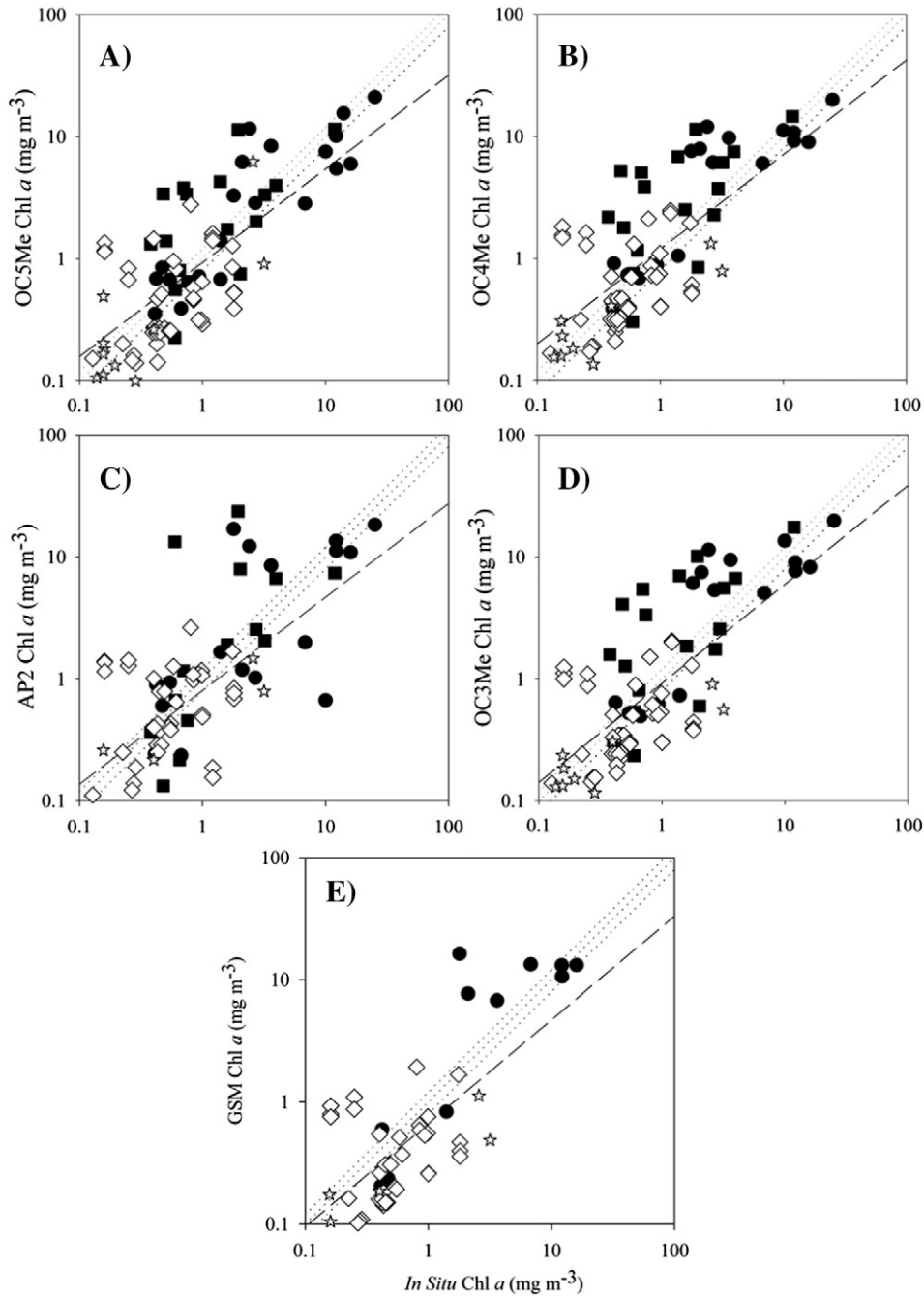


Fig. 3. Comparison of *in situ* Chl *a* and COASTCOLOUR L2 R_{rs} derived Chl *a* for (A.) OC5Me, (B.) OC4Me, (C.) AP2, (D.) OC3Me and (E.) GSM. Faint dotted lines are the 1:1 line, upper and lower 20% quartiles. Dashed line is the regression line. Filled circles are data from the North Sea, filled squares are from the English Channel, open diamonds are from the Portuguese Shelf, open stars are the Mediterranean Sea.

by inverting the semi-analytic model. $nL_w(560)$ is used in clear or moderately turbid waters (where non-algal SPM $< 4 \text{ g m}^{-3}$) and $nL_w(670)$ for highly turbid waters (Gohin, 2011).

2.7. Algorithm performance

The performance of the Chl *a* algorithms was assessed in four ways: (1.) using 482 coincident *in situ* measurements of $R_{rs}(\lambda)$ and Chl *a*; the *in situ* $R_{rs}(\lambda)$ were used to run the algorithms and the resulting Chl *a* values were compared against the *in situ* Chl *a* (Fig. 2, Table 2). (2.) A data base of 594 *in situ* HPLC and fluorometric Chl *a* (which includes the 482 coincident *in situ* $R_{rs}(\lambda)$ and Chl *a*, plus additional Chl *a*

measurements) comprising data from the Celtic Sea (PML), Eastern English Channel (IFREMER), Iberian Peninsula (CIMA & MARE), North Sea (RBINS & PML), Mediterranean Sea (LOV) and Western English Channel (PML) (Fig. 1B), were compared against Chl *a* match-ups from MERIS overpasses using COASTCOLOUR (Fig. 3, Table 2) and MEGS 8.0 (Fig. 4, Table 2) data. MERIS match-ups $\pm 0.5 \text{ h}$ between *in situ* sampling and MERIS data from a 3×3 pixel array around the sampling station were extracted to compare against *in situ* Chl *a* following the procedures outlined in (Bailey and Werdell, 2006). The Chl *a* algorithms yielded a varying number of match-ups depending on the input $R_{rs}(\lambda)$ used and the associated quality flags raised. To facilitate statistical comparison, common returns for OC5Me, OC4Me, OC3Me & AP2 were used. (3.)

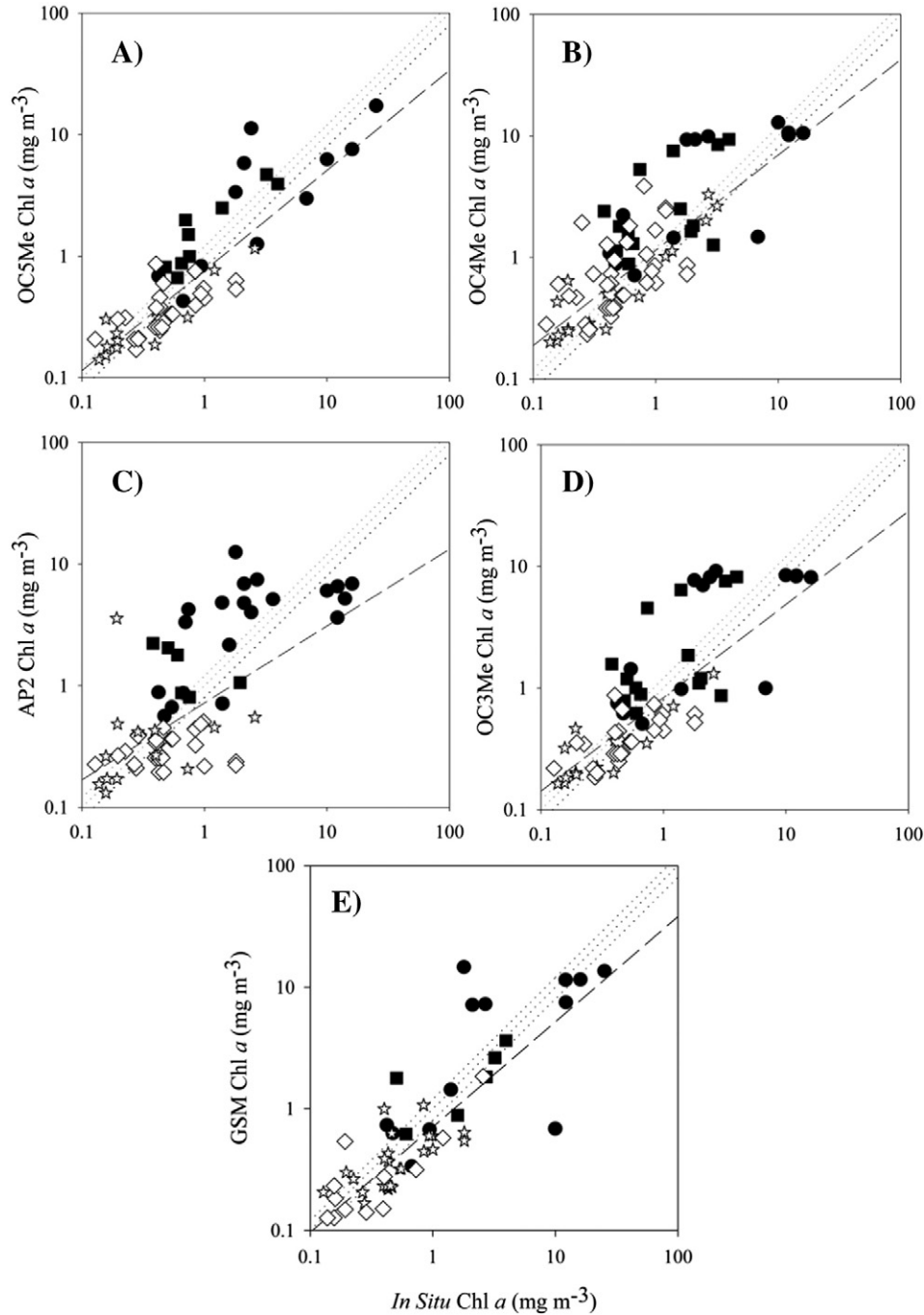


Fig. 4. Comparison of *in situ* Chl *a* and MEGS8.0 L2 R_{rs} derived Chl *a* for (A.) OC5Me, (B.) OC4Me, (C.) AP2, (D.) OC3Me and (E.) GSM. Faint dotted lines are the 1:1 line, upper and lower 20% quartiles. Dashed line is the regression line. Filled circles are data from the North Sea, filled squares are from the English Channel, open diamonds are from the Portuguese Shelf, open stars are the Mediterranean Sea.

MERIS MEGS 8.0 Chl *a* values from AP2, OC4Me and OC5Me were compared along transects in the North Sea from the Thames Estuary in the UK (51.47°N, 0.59°E) to the Schelde Estuary in the Netherlands (51.49°N, 3.31°E) and from the Schelde Estuary (51.49°N, 4.13°E) to

The Wash in the UK (52.67°N, 2.01°E). Chl *a* values were extracted at every 10 km along each transect from daily images between March and September from 2003 to 2007 (Fig. 5). (4.) Monthly composite images from AP2, OC4Me, GSM and OC5Me using MERIS MEGS 8.0 Chl *a*

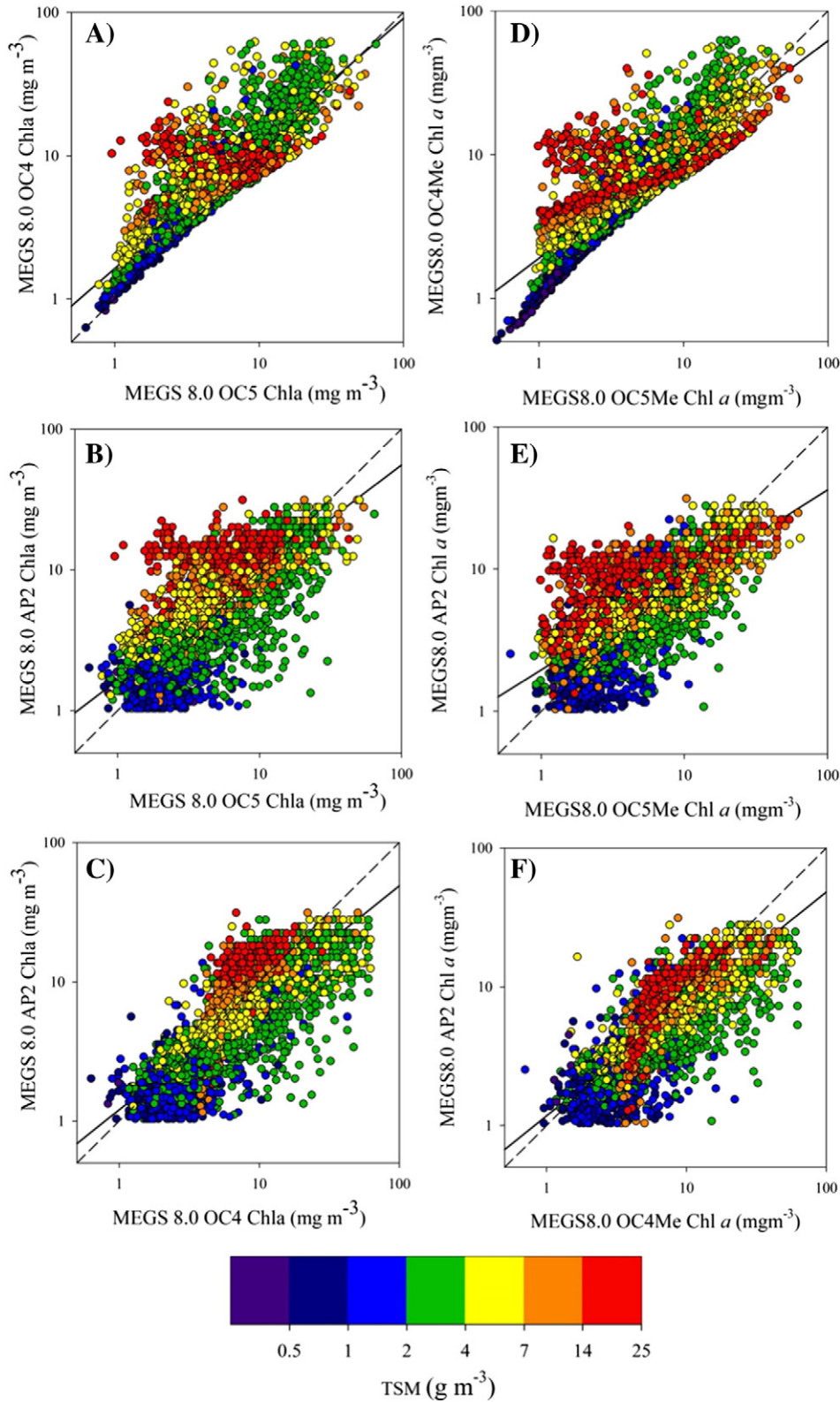


Fig. 5. Comparison of OC5Me, OC4Me and AP2 processed using MEGS8.0 R_{rs} along transects from the Schelde estuary to The Wash; (A.) OC4Me versus OC5 Me, (B.) AP2 versus OC5Me and (C.) AP2 versus OC4Me; and from the Thames to the Schelde estuaries; (D.) OC4Me versus OC5Me, (E.) AP2 versus OC5Me and (F.) AP2 versus OC4Me. Coloured circles indicate non-algal SPM concentration ($g\ m^{-3}$) estimated using the Gohin (2011) algorithm.

were compared for the months of April 2010 and July 2011 (Figs. 6, 7). Additional data were extracted from a transect from the UK coast in the Western English Channel (50.25°N, 4.21°W) to the River Seine on the French coast (49.42°N, 0.14°E) at every 10 km to evaluate spatial differences in algorithm performance for these images (Fig. 8).

To assess algorithm performance, the following statistical metrics were used: The mean (M), standard deviation (S), and \log_{10} -root-mean square (\log_{10} RMS) of the difference error (r) between measured and MERIS Chl *a* at each station following the methods described in Campbell et al. (2002). The geometric mean and one-sigma range of the inverse transformed ratio between satellite and measured values are given by M (F_{med}), M-S (F_{min}), M + S (F_{max}) and were used as algorithm performance indices. The relative percentage difference (RPD) was calculated following Antoine et al. (2008). The uncertainty in MEGS8.0 and COASTCOLOUR $R_{rs}(\lambda)$ were assessed against *in situ* $R_{rs}(\lambda)$ at match-up stations. The relative difference between *in situ* and satellite $R_{rs}(\lambda)$ was calculated following Hu et al. (2013) and expressed as percentage uncertainty over the range in *in situ* Chl *a* and SPM (Fig. 11).

3. Results

3.1. Accuracy assessment of ocean colour algorithms for MERIS in North West European waters using *in situ* $R_{rs}(\lambda)$

Firstly we assessed the accuracy of the algorithms using *in situ* R_{rs} to calculate Chl *a* (Fig. 2, Table 2) measured at the stations given in Fig. 1A. The majority of *in situ* R_{rs} used to compute algorithm Chl *a* values were from the North Sea, which comprised 80% of the dataset. *In situ* R_{rs} from the Portuguese coast and Mediterranean Sea constituted 20% of the data set. The log-RMS, RMS-E bias and random error were smallest for

OC4Me and OC3Me. The percentage variance explained (R^2) was greatest for OC4Me and OC3Me had the lowest RPD, which was within 24 and 9% of *in situ* Chl *a*, respectively (Fig. 2B, D, Table 2). The linear regression slope and F_{min} for GSM were closest to 1, indicating that the GSM is more accurate at the lower range of Chl *a* values when using *in situ* R_{rs} . The GSM however, had the highest log-RMS, RMS-E, F_{max} , F_{med} and intercept (Fig. 2C, Table 2), which suggests that it may not be suitable for these coastal waters. By comparison, OC5 had the lowest intercept and F_{max} closest to 1, indicating that it is more accurate at the higher range Chl *a* values when using these *in situ* R_{rs} data (Fig. 2A, Table 2). The processing of AP2 using *in situ* R_{rs} was not possible as the atmospheric correction is embedded within the standard AP2 processor such that any other input source of R_{rs} other than MERIS R_{rs} , is not feasible.

3.2. Accuracy assessment of MERIS MEGS8.0 and COASTCOLOUR Chl *a* products in North West European waters

Using COASTCOLOUR L2 R_{rs} to test algorithm accuracy (Fig. 3, Table 2), OC4Me had the lowest RPD and r and F_{max} closest to 1. The GSM had the lowest intercept and RMS-E and slope closest to 1, though fewer match-ups were available using the GSM, since the algorithm did not converge at some stations in the North Sea and at all of the English Channel stations (Table 2). OC5Me had the lowest \log_{10} -RMS, M and S and R^2 and F_{min} closest to 1. OC5Me with COASTCOLOUR L2 R_{rs} had a similarly low intercept, random error and RMS-E and F_{med} compared to the GSM and AP2. Of the algorithms tested using COASTCOLOUR L2 R_{rs} , AP2 was the least accurate (Table 2, Fig. 3).

Similarly using MEGS8.0 L2 R_{rs} , OC4Me had the lowest RPD and OC3Me had the lowest random error (Table 2, Fig. 4B, D). OC5Me had the lowest \log_{10} -RMS, RMS-E, intercept and bias, the highest R^2 and

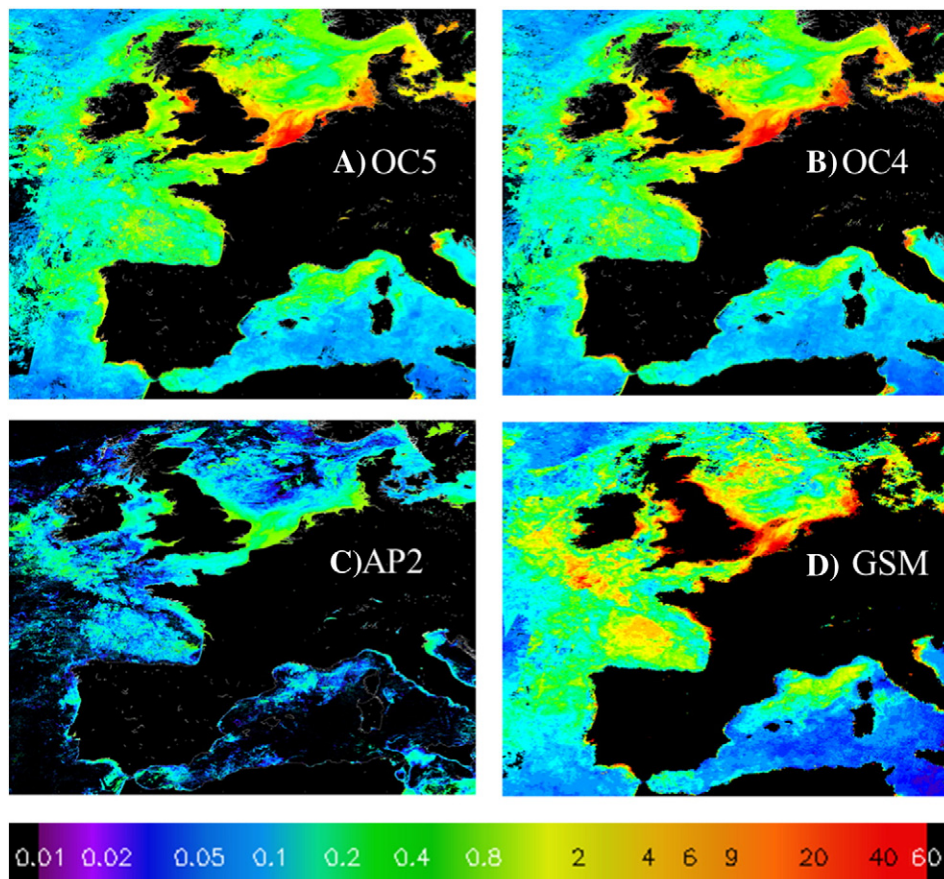


Fig. 6. Comparison of MERIS MEGS8.0 Chl *a* monthly composites for April 2010 using (A.) OC5Me, (B.) OC4Me, (C.) AP2, (D.) GSM.

the slope, r , and F_{\min} closest to 1 and was the most accurate algorithm in 7 out of the 12 statistical tests performed (Table 2, Fig. 4A). GSM also had F_{\min} close to 1 whereas F_{\max} was high, indicating that it is more accurate at lower rather than at higher Chl a concentrations (Table 2, Fig. 4E). For GSM with MEGS8.0 L2 R_{rs} , there were only four data points that exhibited high scatter from the 1:1 line, and in the absence of these, this algorithm would have been accurate for these coastal waters. Where the other algorithms returned data, the GSM did not converge at some stations in the English Channel however, questioning its suitability for providing contiguous satellite imagery. AP2 was the least accurate across all statistical tests (Table 2, Fig. 4C). Of the MEGS8.0 and COASTCOLOUR match-ups, 17% were from the North Sea, 13% from the English Channel, 55% were from the Portuguese coast and 15% from the Mediterranean. Comparing MEGS8.0 and COASTCOLOUR OC5Me Chl a against *in situ* Chl a and using the same number of match-ups (figure not shown), MEGS8.0 OC5Me had a lower bias and log-RMS and a higher percentage variance explained compared to COASTCOLOUR OC5Me, indicating that OC5Me Chl a is slightly more accurate using MEGS8.0 compared to COASTCOLOUR R_{rs} (Table 2, Figs. 3A, 4A).

3.3. Spatial comparison between MERIS ocean colour algorithms in the North Sea

A comparison of Chl a from MERIS algorithms along two transects from the Schelde Estuary and to the Wash on the SE UK coast and from the Schelde to the River Thames estuary across the Southern Bight of the North Sea using data from January 2003 to December 2007 (Figs. 1C, 5). For these two transects over the 6 yr period, there were 68,332 data points; for coincident OC5Me, OC4Me AP2 and non-algal SPM data there were 7776 data. Comparison between OC4Me and OC5Me explained 58% of the variance in the data with a slope

close to 1, but a high intercept (Table 3). At Chl $a > 1 \text{ mg m}^{-3}$ OC5Me, OC4Me tended to over-estimate Chl a which were $\sim 10 \text{ mg m}^{-3}$ (Fig. 5A, D). At OC5Me Chl a between 1 and 10 mg m^{-3} , the over-estimate in OC4Me Chl a was related to high non-algal SPM $> 14 \text{ g m}^{-3}$. At OC5Me Chl $a > 10 \text{ mg m}^{-3}$, the over-estimate in OC4Me was not related to SPM which was between 2 and 8 g m^{-3} . The offset between OC5Me and OC4Me was worse on the Thames compared to the Wash transects due to an increase in SPM adjacent to the Thames estuary. For OC5Me and AP2, 51% of the variance between the two algorithms was explained, the slope was low (~ 0.58), the intercept was high (Table 3) and there was a large offset between the algorithms at both low and high Chl a concentrations (Fig. 5B, E). There was a large error in AP2 Chl a between 1 and 20 mg m^{-3} at both high ($> 14 \text{ g m}^{-3}$) and low ($2\text{--}3 \text{ g m}^{-3}$) non-algal SPM. The slope was lower on the Thames transect but the scatter was higher on the Wash transect (Fig. 5B, E, Table 3). Above 10 mg m^{-3} Chl a , the scatter between OC5Me and AP2 was reduced. Similarly, for OC4Me and AP2, 45% of the variance was explained, the slope was low and the intercept was high (Table 3) and there was a large scatter between the two algorithms over the entire OC4 Chl a range from $0.6\text{--}65.0 \text{ mg m}^{-3}$. The offset and scatter between the algorithms was similar on both transects (Fig. 5C, F). The spatial and temporal differences between the algorithms are illustrated in satellite images during two different periods in spring (April 2010; Fig. 6) and summer (July 2011; Fig. 7). In coastal waters during both periods, OC5Me Chl $a < \text{OC4Me Chl } a$, whereas in open ocean waters OC5Me and OC4Me were similar (Fig. 6) except in July 2011 when OC4Me $> \text{OC5Me}$ (Fig. 7). By contrast, AP2 exhibited the lowest Chl a in coastal regions of the North Sea and English Channel and did not yield data at all in some coastal and shelf areas (Figs. 6C, 7C). For GSM, in the coastal regions off Denmark, Holland, Belgium, northern France and the UK, Chl a was higher than for OC5Me and OC4Me (Figs. 6D,

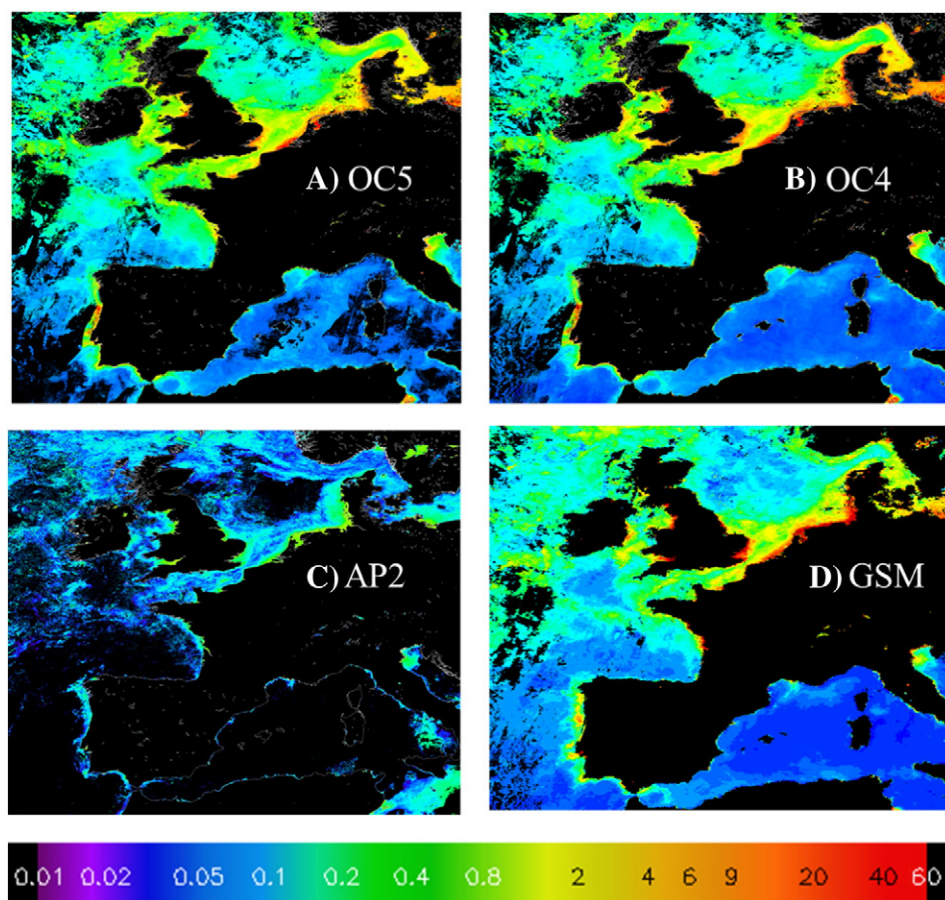


Fig. 7. Comparison of MERIS MEGS8.0 Chl a monthly composites of July 2011 using (A.) OC5Me, (B.) OC4Me, (C.) AP2, (D.) GSM.

7D). In some offshore areas of the North Sea and Celtic Sea off SW England and southern Ireland, GSM Chl *a* was lower, but in spring in the Bay of Biscay, off the west Irish Coast and central North Sea, GSM Chl *a* was also higher. To highlight the spatial variability in the satellite images between the algorithms, Chl *a* were extracted from the April 2010 images (Fig. 6) at every 10 km along transects from the Wash to the Schelde Estuary and from the English Channel to the Seine Estuary on the French coast (Fig. 1C). At the ends of the transects near to the UK coast, OC4Me Chl *a* was 2 to 6 times higher than OC5Me (Fig. 8A, B, E, F). AP2 under-estimated Chl *a* across the entire transect (Fig. 8C, G). GSM generally over-estimated Chl *a* but the spatial trends were erratic with large oscillations between high and low Chl *a* values over small spatial scales (Fig. 8D, H).

4. Discussion

4.1. MERIS ocean colour algorithm accuracy assessment

The principal objective of this study was to assess the performance of Chl *a* algorithms, that are widely available to end users, with two different MERIS AC processors in North West European waters. *In situ* R_{rs} were firstly used to test the algorithms in the absence of the MERIS AC model to indicate potential errors in the Chl *a* algorithm alone rather than combined errors arising from the input MERIS R_{rs} and the Chl *a* algorithm. Using *in situ* R_{rs} , OC5Me, OC4Me and OC3Me had a similar slope, but OC5Me showed a higher scatter, illustrated by the higher S and M values compared to OC4Me and OC3Me (Table 2). This suggests that with *in situ* R_{rs} , OC4Me and OC3Me are the most accurate algorithms for the NW European region (Fig. 2B, D). The slope for GSM was closer to 1, but the scatter from 0.5 to 20 mg m⁻³ was high (Fig. 2C) and this algorithm was the least accurate using *in situ* R_{rs} . It was

not possible to run AP2 with just *in situ* R_{rs} , since the AC is an integral part of the AP2 NN processor and is not available without the AC component of the model.

Secondly, both COASTCOLOUR and MEGS8.0 R_{rs} were used to test the algorithms. We found that AP2 was the least accurate in these regions (Figs. 3C, 4C); AP2 had a large bias using both processors at low and high Chl *a* concentrations, which is reflected in the high log₁₀-RMS, S, F_{min} and F_{max} values (Table 2). When AP2 was processed as MEGS8.0 monthly composite images for European coastal regions, there were a large number of pixel drop-outs caused by none convergence in the NN (Figs. 6C, 7C). In addition, AP2 Chl *a* values in both coastal and open ocean areas of the North Sea and Atlantic Ocean were significantly lower than those obtained from OC4Me and OC5Me (Figs. 6, 7, 8), which is reflected by the tendency to under-estimate Chl *a* shown in Figs. 3C, 4C.

The AP2 NN has been shown to be accurate in the North Sea (Doerffer and Schiller, 2007), but is not as accurate in the Adriatic (Zibordi et al., 2006a), Baltic (Attila et al., 2013), Mediterranean (Antoine et al., 2008) Seas, South African coastal and shelf waters (Smith et al., 2013), Siberian Arctic coastal and shelf waters (Heim et al., 2014), US coastal regions (Mishra and Mishra, 2012) and tropical coastal regions (Ambarwulan et al., 2010), especially when compared against standard MODIS-Aqua or SeaWiFS ocean colour products. In a number of studies the MERIS AP2 NN has been modified to account for the regional variation in IOPs which has improved its performance (Attila et al., 2013; Smith et al., 2013). Alternative algorithms have been proposed for MERIS that have proven to be more accurate for the North Sea (Tilstone et al., 2012; Van der Woerd and Pasterkamp, 2008) and Iberian Peninsula (Gonzalez Vilas et al., 2011; Sa et al., 2015), but these are not widely available to end users. There is a tendency for MERIS AP2 in some coastal waters to under-estimate Chl *a* in the

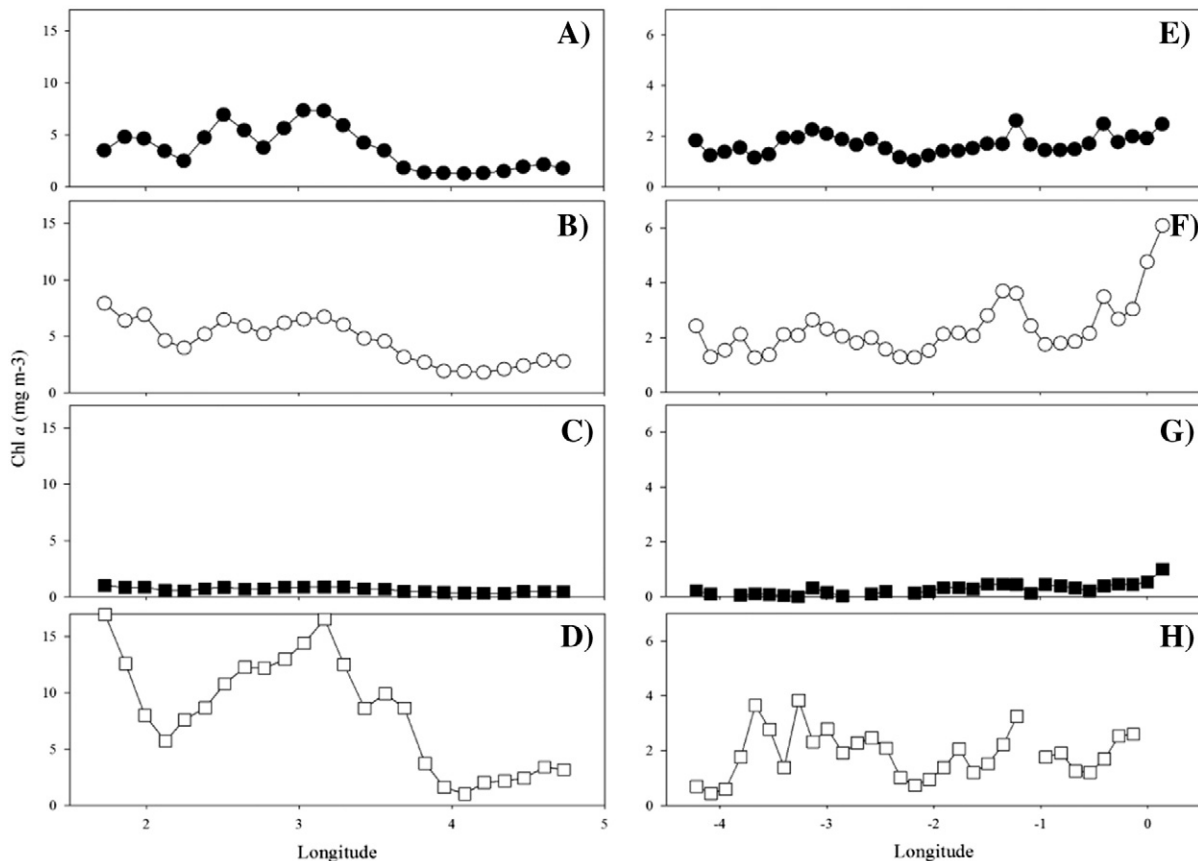


Fig. 8. Comparison of MERIS MEGS8.0 Chl *a* from the monthly composite of April 2010 (given in Fig. 6) along Longitudinal transects from the Wash to the Schelde for (A.) OC5Me, (B.) OC4Me, (C.) AP2, (D.) GSM and from the western English Channel to the River Seine for (E.) OC5Me, (F.) OC4Me, (G.) AP2, (H.) GSM. The location of the transects are given in Fig. 1C.

Table 3

Percentage variance explained (R^2), intercept and slope from linear regression between OC5Me, OC4Me and AP2 Chl a along transects from the Schelde estuary to The Wash and the Thames to the Schelde estuaries (see Fig. 1C for details of transects).

	R^2	Slope	Intercept	R^2	Slope	Intercept
	The wash-schelde estuary N = 3734			Thames-schelde estuaries N = 4042		
OC5Me v OC4Me	0.60	1.13	1.36	0.57	0.87	2.27
OC5Me v AP2	0.49	0.64	2.77	0.50	0.53	3.00
OC4Me v AP2	0.44	0.42	3.09	0.47	0.43	2.85

range $< 1 \text{ mg m}^{-3}$ and to over-estimate Chl a at values $> 6 \text{ mg m}^{-3}$ (Tilstone et al., 2012). Knowing this bias in the algorithm, MERIS AP2 has under gone several rounds of re-calibration, which has resulted in

re-processing the MERIS archive (Bourg et al., 2011). One of the problems experienced with the AP2 N is that adding further training data can lead to 'over-training' which offers the algorithm further multiple possible solutions to retrieving IOPs, which may not necessarily result in accurate IOP spectra (IOCCG, 2006). From our analyses, the scatter plots indicated that COASTCOLOUR AP2 under- and over-estimates Chl a in all regions tested and that MEGS8.0 AP2 under- and over-estimates Chl a in the North Sea and English Channel, and under-estimates in the Mediterranean Sea and off the Portuguese coast (Figs. 3C, 4C). Analysis of the satellite imagery indicated that AP2 consistently under-estimated Chl a in all regions (Figs. 6C, 7C, 8C, G).

By comparison, OC4Me was designed for global applications over optically deep ocean waters (Morel and Antoine, 2011). The retrieval accuracy of Chl a by satellite ocean colour sensors is expected to be

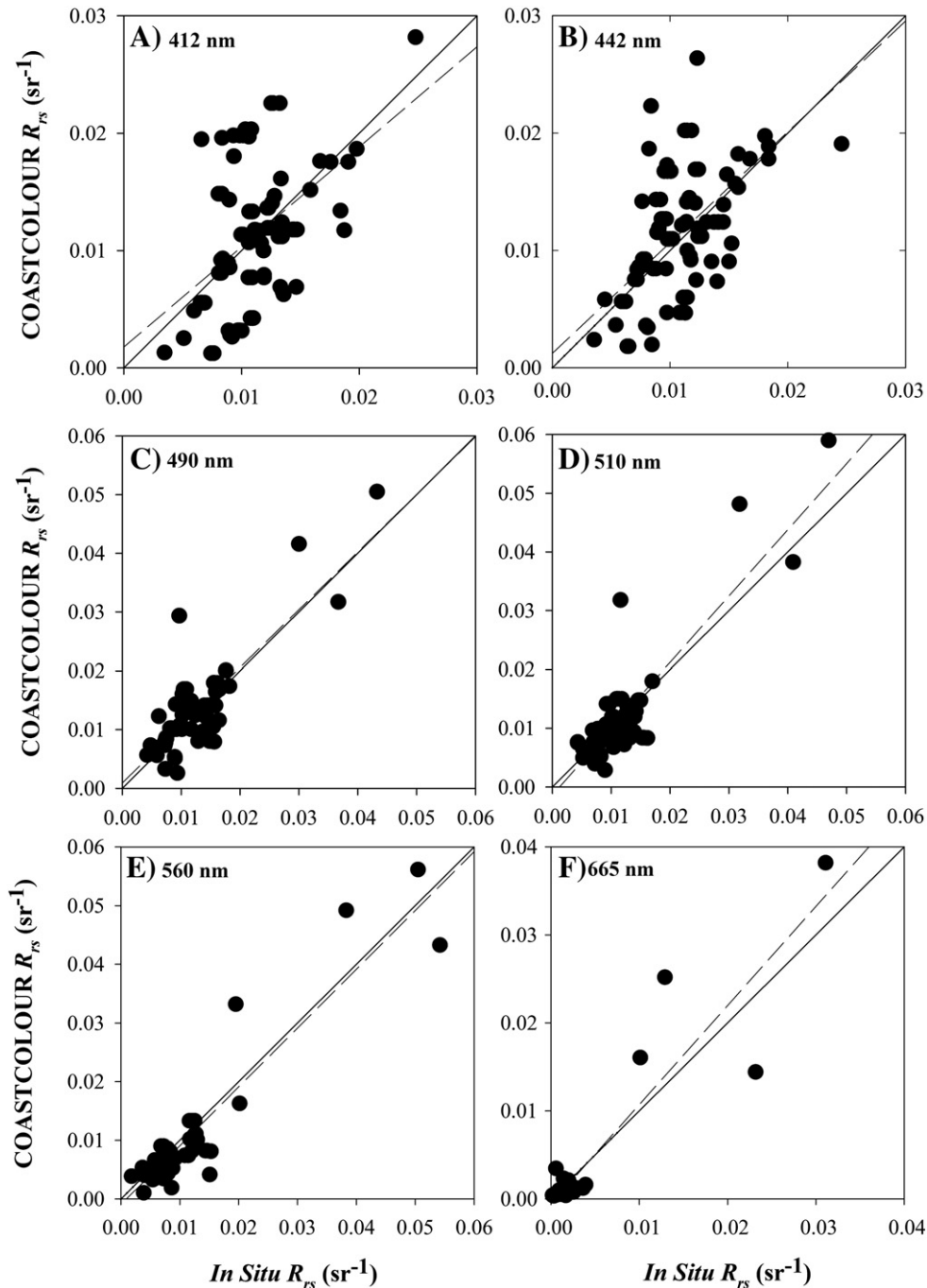


Fig. 9. Scatter plots of *in situ* $R_{rs}(\lambda)$ against COASTCOLOUR $R_{rs}(\lambda)$ for (A.) 412, (B.) 442, (C.) 490, (D.) 510, (E.) 560 and (F.) 665 nm. Solid line is the 1:1; dashed line is the regression line.

within $\pm 35\%$ in oceanic waters (Bailey and Werdell, 2006; Blondeau-Patissier et al., 2014). This may not always be the case (Hu et al., 2000; Moore et al., 2009) and therefore may not be sufficient for the purpose of monitoring phytoplankton biomass in the coastal zone. In the coastal waters tested, OC4Me consistently over-estimated Chl *a* at values $> 1 \text{ mg m}^{-3}$ (Figs. 3B, 4B), though for the few match-ups with Chl *a* $> 10 \text{ mg m}^{-3}$, the retrieval accuracy improved. The use of these blue: blue-green band-ratios often leads to erroneous retrievals in coastal waters, where the optical complexity is highly variable (Blondeau-Patissier et al., 2004). The blue band can be affected by both the absorption of CDOM and scattering from SPM (Dierssen, 2010) and in turbid waters the blue-green band can be related to SPM more than to phytoplankton. Absorbing aerosols and the a_{CDOM} can

propagate as negative nL_w at short wavelengths, which can result in an over-estimate of Chl *a*, as observed in coastal regions of the Bay of Bengal (Tilstone et al., 2011), United States (Cannizzaro et al., 2013; Le et al., 2013a), the Bering Sea (Naik et al., 2013), South-East Asia (Ahn and Shanmugam, 2006), the North Sea and Kattegat (Jorgensen, 2004). OC4 is often applied indiscriminately in some coastal waters (Dupouy et al., 2010), without a proper understanding of the potential errors that can be incurred in these waters. For the coastal and shelf environments tested in this study however, the relative difference between the *in situ* measurements and OC4Me over the entire Chl *a* range was $\sim 25\%$, and therefore within the accepted tolerance for Case 1 waters. OC4Me tended to over-estimate Chl *a* in coastal areas of the North Sea, English Channel and Portuguese coast (Figs. 3B, 4B, 8B, F).

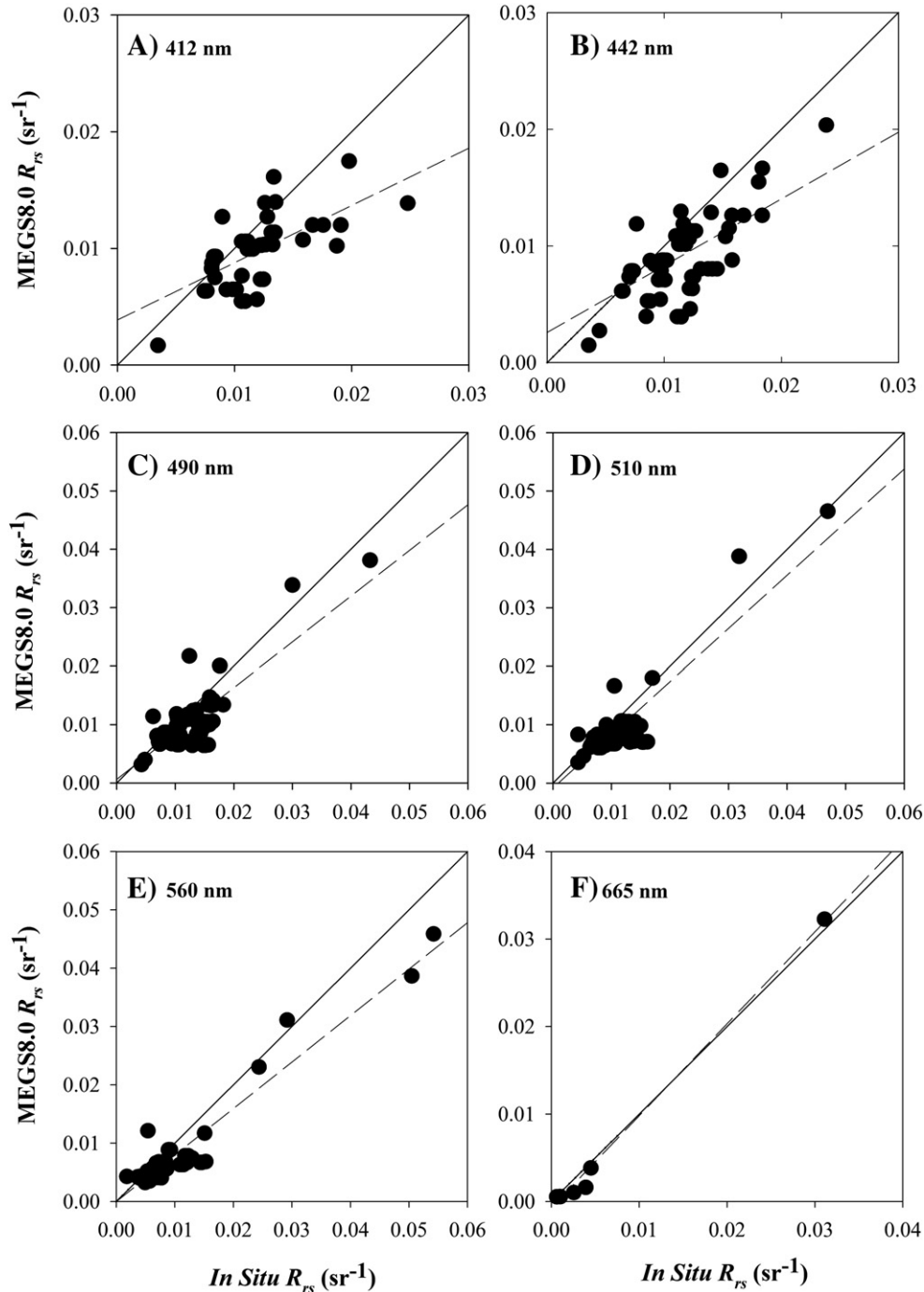


Fig. 10. Scatter plots of *in situ* $R_{rs}(\lambda)$ against MEGS8.0 $R_{rs}(\lambda)$ for (A.) 412, (B.) 442, (C.) 490, (D.) 510, (E.) 560 and (F.) 665 nm. Solid line is the 1:1; dashed line is the regression line.

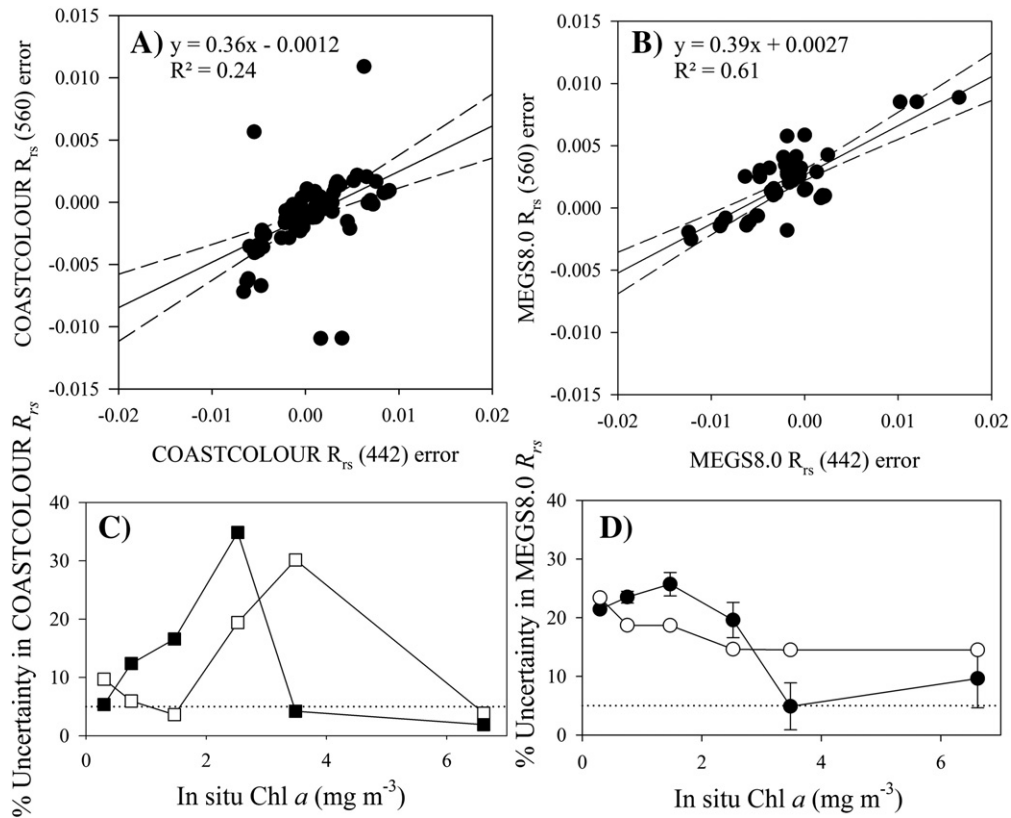


Fig. 11. Spectral relationships between R_{rs} errors for (A.) MEGS8.0 $R_{rs}(442)$ and $R_{rs}(560)$ and (B.) COASTCOLOUR $R_{rs}(442)$ and $R_{rs}(560)$. The errors are determined as the relative difference between MERIS and *in situ* R_{rs} . Percentage uncertainty in (C.) COASTCOLOUR $R_{rs}(442)$ (open squares) and $R_{rs}(560)$ (filled squares) and (D.) MEGS8.0 $R_{rs}(442)$ (open circles) and $R_{rs}(560)$ (filled circles) as a function of Chl *a*. In (A.) and (B.), the solid line is the regression and the dashed lines are the 95% confidence intervals. In (C.) and (D.), the dashed line is the 5% uncertainty limit.

The performance of OC3Me was similar to OC4Me, with the advantage of omitting the $R_{rs}(412)$ band where the error is greatest (see Section 4.2). For OC3Me, of the ~90 satellite-*in situ* Chl *a* match ups obtained with MEGS8.0 (Fig. 4D), 83% of these used the $R_{rs}(490): R_{rs}(560)$ ratio. The $R_{rs}(442): R_{rs}(560)$ ratio was used to calculate Chl *a* at stations predominantly from the Portuguese coast and Mediterranean Sea, which relate to those points below the 1:1 in Fig. 4D, where the under-estimate in $R_{rs}(442)$ (see Section 4.2) led to lower Chl *a* concentrations. The high scatter in OC3Me Chl *a* in the North Sea were predominantly at stations in the southern North Sea off the coast of France, principally due to an under-estimate in $R_{rs}(490)$ (see Section 4.2), which led to higher Chl *a* values (Figs. 3D & 4D).

The GSM algorithm had a slope close to 1 and a low intercept though there was a very high scatter for some of the match-up points, which resulted in high log-RMS and RPD (Figs. 3E, 4E). This was also reflected in the composite images from April 2010 (Fig. 6) and July 2011 (Fig. 7), which showed that GSM Chl *a* was consistently higher in coastal regions and both lower and higher offshore compared with OC5Me and OC4Me. The GSM Chl *a* extracted from transects in April 2010 in the North Sea and English Channel (Fig. 8D, H) indicate that the outliers in the scatter plot represent large areas of the image where GSM consistently over-estimate Chl *a* (Fig. 8). The GSM is semi-analytical and was calibrated using the SeaBAM dataset (Maritorena et al., 2002) and like OC4, was originally developed for Case 1 waters. The GSM uses $R_{rs}(442)$ to

Table 4

Performance indices for relative errors between *in situ* and COASTCOLOUR and MEGS8.0 $R_{rs}(\lambda)$ at visible wavebands. Percentage variance explained (R^2), intercept and slope and log-difference errors in measured and satellite Chl *a* ratio as Mean (*M*), Standard deviation (*S*) and root-mean square (Log_{10} -RMS).

λ	N	R^2	Slope	Intercept	R	RPD	Log_{10} -RMS	M	S	F_{med}	F_{max}	F_{min}	RMS-E
COASTCOLOUR													
412	98	0.53	1.14	0.004	1.02	0.67	0.25	0.05	0.25	1.11	1.98	0.62	0.005
442	103	0.72	1.02	-0.003	1.00	5.33	0.19	0.01	0.20	1.02	1.60	0.65	0.004
490	101	0.65	0.98	0.0009	0.99	0.19	0.15	-0.01	0.15	0.97	1.36	0.69	0.004
510	101	0.61	1.23	-0.002	1.01	16.98	0.13	0.01	0.13	1.03	1.39	0.77	0.004
560	100	0.86	0.86	0.002	1.03	69.06	0.16	0.07	0.14	1.17	1.63	0.84	0.003
665	91	0.76	1.28	0.0006	1.06	132.66	0.26	0.19	0.17	1.54	2.30	1.04	0.001
MEGS8.0													
412	68	0.41	0.49	0.004	1.05	-9.2	0.14	0.09	0.11	1.22	1.58	0.94	0.0005
442	70	0.46	0.54	0.003	1.06	-8.3	0.20	0.11	0.16	1.29	1.87	0.89	0.004
490	94	0.63	0.78	0.0006	1.05	-6.8	0.14	0.09	0.11	1.23	1.58	0.96	0.0007
510	94	0.75	0.91	-0.0008	1.05	2.8	0.14	0.09	0.10	1.23	1.56	0.98	0.001
560	88	0.89	1.12	0.0009	1.05	32.5	0.15	0.11	0.10	1.27	1.62	1.00	0.0006
665	18	0.99	1.05	-0.0007	1.07	84.1	0.25	0.18	0.14	1.53	2.11	1.11	0.0003

partition the absorption $a_{dg}(442)$ and $b_{bp}(442)$ and is calibrated using a simulated annealing procedure to retrieve $a_{ph}^*(\lambda)$ at 412, 443, 488, 530 and 555 nm. It solves $a_{ph}(\lambda)$ in the presence of SPM and $a_{CDOM}(\lambda)$ by using a constant exponential slope of CDOM (S_{CDOM}) of $0.0206 \text{ m}^{-1} \text{ nm}^{-1}$, a power-law exponent for particulate backscattering ($\eta = 1.03373$) expressed as a function of Chl a (Morel and Maritorena, 2001), and an optimized a_{ph}^* as a fixed value [$a_{ph}^*(443) = 0.05582$]. This IOP parameterisation may not be appropriate for these European coastal waters; S_{CDOM} used for calibrating the GSM is too high for the North Sea; S_{CDOM} of 0.0101 and $0.0232 \text{ m}^{-1} \text{ nm}^{-1}$ have been reported for this area (Astoreca et al., 2009) and therefore $0.0206 \text{ m}^{-1} \text{ nm}^{-1}$ is towards the upper limit for these waters. Similarly the assigned $a_{ph}^*(\lambda)$ is too low for the Celtic Sea and the English Channel which have mean values between 0.07 & 0.09 mg m^{-2} , respectively (Tilstone et al., 2012). With MEGS 8.0, the under-estimate in $R_{rs}(442)$ (Fig. 10B) will propagate into the partitioning of $a_{dg}(442)$ and $b_{bp}(442)$ and then to estimating $a_{ph}^*(442)$, which resulted in an over-estimate in Chl a (Fig. 4E). This was observed over the entire region particularly in coastal and shelf regions of the North Sea and English Channel in spring (Figs. 6D, 8D, H).

OC5 was initially developed to provide realistic maps of Chl a for the turbid waters of the English Channel and the Bay of Biscay for validating biogeochemical model outputs (Menesguen et al., 2007). The first SeaWiFS OC4v4 products showed blooms in January 1998 in the English Channel, which never occur and therefore prevented the use of these products being used for model assessment. High SeaWiFS R_{rs} in blue and green bands for this region during winter, due to an increase in scattering from SPM, caused the over-estimate OC4v4 Chl a . The rationale for developing OC5 was therefore to provide more accurate estimates of Chl a in turbid waters and to be as close as possible to OC4 Chl a in Case 1 waters. By construction, OC5 Chl a is always lower than OC4 Chl a , which makes its application in Case 1 waters less reliable (Marrec et al., 2015). In the shelf waters of the Celtic and North Seas for example, OC5Me was similar, but slightly lower than OC4Me (e.g. Figs. 6, 7, 8). Since OC5 is empirical and is calibrated using level 2 $R_{rs}(\lambda)$, when data from a specific satellite sensor or mission, reprocessing may be required before applying OC5 to different data sources (Morozov et al., 2010). This implies that when using $nL_w(\lambda)$ from MODIS, MERIS, merged MERIS-MODIS-SeaWiFS, Sentinel-3 or any variant of these with a different atmospheric correction model or with *in situ* R_{rs} (as illustrated in Fig. 2A), OC5 would require a full re-parameterization. The necessity to re-parameterize OC5Me when using a different $nL_w(\lambda)$ source is illustrated in Fig. 2. Since the LUT for OC5Me was parameterized using MERIS $R_{rs}(\lambda)$, using *in situ* $R_{rs}(\lambda)$ to run the algorithm resulted in a high scatter at Chl $a > 1 \text{ mg m}^{-3}$ in the North Sea (Fig. 2A) and was the third most accurate algorithm after OC3Me and OC4Me. Despite these limitations, OC5 has proven to be accurate in a range of different coastal waters including the Ganges Delta in the Bay of Bengal (Tilstone et al., 2011), the Bay of Biscay (Novoa et al., 2012) and the Ligurian and Tyrrhenian Seas (Lapucci et al., 2012). In NW European waters OC5Me was the most accurate algorithm using both COASTCOLOUR and MEGS8.0 $R_{rs}(\lambda)$ over the range of 0.1 to 35.0 mg m^{-3} Chl a . OC5Me was parameterised using the OC4Me maximum band ratio and $nL_w(412)$ and $nL_w(560)$ bands from MERIS data in the English Channel and Bay of Biscay, where Chl a is 0.05 to 15 mg m^{-3} , SPM are 0.1 to 10 g m^{-3} (Gohin et al., 2002) and $a_{CDOM}(375)$ is 0.02– 1.76 m^{-1} (Vantrepotte et al., 2007). The range in Chl a , SPM and $a_{dg}(442)$ in North Sea, Mediterranean Sea, Western English Channel and Portuguese waters was 0.13– 25.13 mg m^{-3} Chl a ; 0.001– 27.02 g m^{-3} SPM; 0.22– 1.80 m^{-1} $a_{dg}(442)$, which covers the OC5 parameterisation range. Since OC5 was originally parameterised using

in situ data from the English Channel and the neighbouring Bay of Biscay, it is perhaps not surprising that OC5Me performed so well. Further testing of OC5Me in other coastal and shelf seas is necessary to assess the applicability of this algorithm globally and whether further parameterisation may be necessary for other water types. Other empirical algorithms such as red: NIR, red: green, fluorescent line height and normalized difference Chl a index algorithms have been shown to be accurate with MERIS R_{rs} in optically complex waters at Chl $a > 10 \text{ mg m}^{-3}$ (Gower et al., 2005; Le et al., 2013b; Mishra and Mishra, 2012; Moses et al., 2012), including freshwater Lakes (Binding et al., 2011; Gilerson et al., 2010). In future studies, it may be interesting to compare these algorithms against OC5Me for coastal and shelf environments.

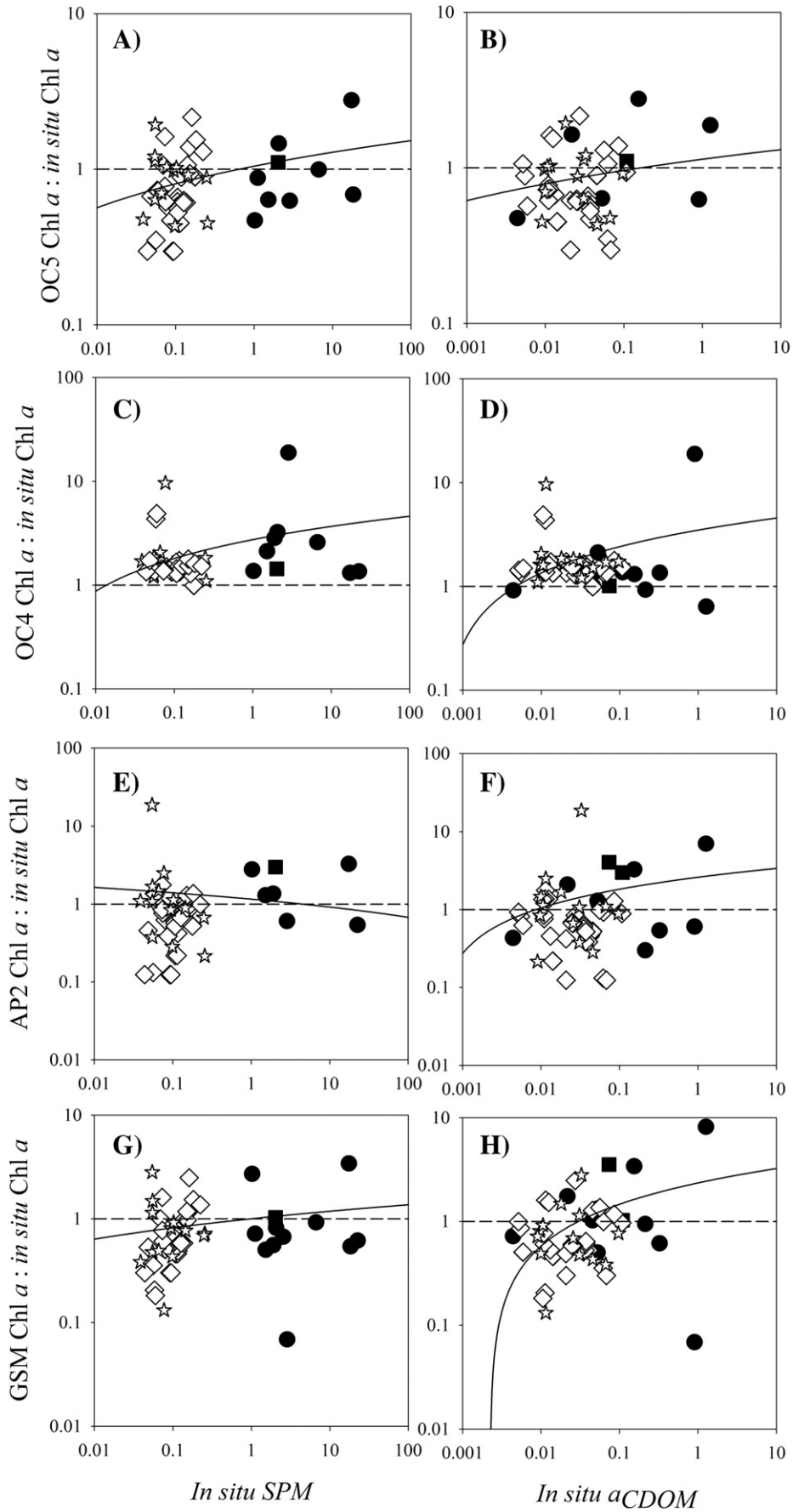
4.2. Causes of differences between MERIS Ocean colour algorithms

To assess the performance of MERIS Chl a algorithms in NW European waters, we firstly addressed the question; what is the accuracy of MERIS $R_{rs}(\lambda)$ derived from different AC processors? MEGS8.0 was designed for use in Case 1 and 2 waters and COASTCOLOUR was designed specifically for use Case 2 waters. We were therefore also able to address the question of whether AC algorithms developed for global water types are more accurate than those developed specifically for Case 2 waters? In Figs. 9 & 10 *in situ* and COASTCOLOUR and MEGS8.0 $R_{rs}(\lambda)$ are compared at MERIS bands. In Fig. 11 the error between $R_{rs}(442)$ and $R_{rs}(560)$ for COASTCOLOUR and MEGS8.0 is compared and the uncertainty in these bands over the range of *in situ* Chl a concentrations is given. The difference between *in situ* and COASTCOLOUR $R_{rs}(442)$ over the match-up dataset was 5% and 8% for MEGS8.0 (Figs. 9, 10; Table 4).

Recent studies based on continuous *in situ* measurement acquisition from towers or buoys have shown that MERIS over-estimates $nL_w(442)$ globally by 44% (Maritorena et al., 2010) and at coastal sites in the Adriatic-Baltic by 39% (Zibordi et al., 2009a; Zibordi et al., 2006b), in the Mediterranean by 36% (Antoine et al., 2008) and in the Skagerrak by 40% (Sorensen et al., 2007), which suggest that both of these MERIS AC processors improve $R_{rs}(442)$ in NW European waters. At least 65% of the stations in the validation data set had SPM $> 3.0 \text{ g m}^{-3}$, which theoretically should not affect these AC processors since both the COASTCOLOUR NN and MEGS8.0 BP AC are optimized for turbid, highly scattering waters. At R_{rs} 490, 510, and 560 nm, the difference between *in situ* and MEGS8.0 and COASTCOLOUR $R_{rs}(\lambda)$ decreased and MEGS8.0 $R_{rs}(560)$ and the NN COASTCOLOUR processor showed a similar accuracy (Figs. 9, 10; Table 4). Though the slope was closer to 1 for COASTCOLOUR.

$R_{rs}(412)$, $R_{rs}(442)$, $R_{rs}(490)$ and $R_{rs}(510)$, compared to MEGS8.0, the scatter was higher which increased the bias and random error (Table 4). Similarly, Goyens et al. (2013) found that MODIS-Aqua with a NN AC model, performed better at blue bands in water masses influenced by SPM, compared to the standard MODIS-Aqua AC. For $R_{rs}(560)$, MEGS8.0 was more accurate than COASTCOLOUR. For $R_{rs}(665)$, the RPD for both MEGS8.0 and COASTCOLOUR in these waters were higher than those reported both globally (~125%), in the Baltic and Adriatic (~47%), the Mediterranean (~70%) and in the Skagerrak (~40%) (Antoine et al., 2008; Zibordi et al., 2006a), though for MEGS8.0 there were fewer points due to a high number of error flags raised. Fig. 11 shows the relationship between the error in R_{rs} (calculated as the difference between MEGS8.0 or COASTCOLOUR $R_{rs}(\lambda)$ and *in situ* $R_{rs}(\lambda)$) at $R_{rs}(560)$ and $R_{rs}(442)$. For MEGS8.0, the error in $R_{rs}(560)$ varied by $> 0.009 \text{ sr}^{-1}$, and for COASTCOLOUR the error was $> 0.015 \text{ sr}^{-1}$, which is ~4 times higher than that reported for the global open ocean (Hu et al., 2013). By comparison, the error in $R_{rs}(442)$ was lower for COASTCOLOUR compared to MEGS8.0, though for both processors this was $> 0.02 \text{ sr}^{-1}$ (Fig. 11A, B). The potential error from atmospheric

Fig. 12. Ratio of MERIS MEGS8.0 Chl a : *in situ* Chl a versus SPM for (A.) OC5Me, (C.) OC4Me, (E.) AP2, (G.) GSM and versus *in situ* $a_{CDOM}(442)$ for (B.) OC5Me, (D.) OC4Me, (F.) AP2, (H.) GSM. Dotted line represents when MERIS Chl $a = in situ$ Chl a . Solid line is line of best fit using a 2 parameter logarithmic fit. Filled circles are data from the North Sea, filled squares are from the English Channel, open diamonds are from the Portuguese Shelf, open stars are the Mediterranean Sea.



correction is estimated as $\pm 0.0006 \text{ Sr}^{-1}$ for $R_{rs}(443)$ in the global open ocean (Gordon et al., 1997). The error in $R_{rs}(\lambda)$ was therefore over an order of magnitude greater than the nominal error due to AC, indicating that other environmental effects (e.g. non-algal SPM backscattering, high absorption by CDOM, sun glint) contribute more to the errors in both MEGS8.0 and COASTCOLOUR $R_{rs}(442)$ (Fig. 11A, B). High uncertainty in $R_{rs}(\lambda)$ may be attributed to errors in the standard aerosol model of optical thickness used (Aznay and Santer, 2009) or failure in the correction at cloud borders (Gomez-Chova et al., 2007). The errors between COASTCOLOUR and MEGS8.0 $R_{rs}(560)$ and $R_{rs}(442)$ were correlated especially for MEGS8.0 (Fig. 11A, B), suggesting that it could be possible to predict the errors between these bands and systematically correct for them. For both COASTCOLOUR and MEGS8.0, the error in the $R_{rs}(560)$ band was less than half of the error at $R_{rs}(442)$ (Fig. 11A, B). The differences reported in Table 4 provide an average uncertainty over the entire $R_{rs}(\lambda)$ match-up data set. The errors in ocean colour Chl *a* however, may not always be uniform (Lee et al., 2010). For example, for SeaWiFS and MODIS-Aqua in oligotrophic regions of the Atlantic and Pacific Oceans over a Chl *a* range of 0.05 to 0.2 mg m^{-3} , Hu et al. (2013) reported an absolute accuracy of <5% for $R_{rs}(\lambda)$ at blue bands for repeat satellite passes. For green and red bands the accuracy was >5% and the uncertainty in both SeaWiFS and MODIS-Aqua $R_{rs}(\lambda)$ tended to increase with increasing Chl *a*. From MERIS match-ups with *in situ* $R_{rs}(\lambda)$, we found that the uncertainty in MEGS8.0 $R_{rs}(442)$ (~17%) was slightly higher than COASTCOLOUR (~12%) over a Chl *a* range of 0.3 to 7 mg m^{-3} (Fig. 11D). At specific Chl *a* concentrations however, there were large variations between the processors. Whilst the uncertainty for MEGS8.0 $R_{rs}(442)$ was fairly constant with increasing Chl *a*, for COASTCOLOUR $R_{rs}(442)$ the uncertainty was low from 0.3 to 1.5 mg m^{-3} and at 6.5 mg m^{-3} , but then increased sharply between 2 and 4 mg m^{-3} Chl *a* (Fig. 11D), reflecting the high scatter at the mid to upper range of $R_{rs}(442)$ values (Fig. 9B). For $R_{rs}(560)$, both MEGS8.0 and COASTCOLOUR processors exhibited a similar pattern, with low uncertainty at low and high Chl *a* concentrations, which for COASTCOLOUR were within the mission goal of absolute accuracy of <5% (Fig. 11D). By comparison, the uncertainty in MEGS8.0 $R_{rs}(560)$ was ~20% from 0 to 2.5 mg m^{-3} and <5% at Chl *a* > 2.5 mg m^{-3} . The net effect of these differences between AC processors is that when they are applied to a range of Chl *a* algorithms, MEGS8.0 proved to be slightly more accurate than COASTCOLOUR. Though the majority of the match-ups points were close to the coast (Fig. 1B), they potentially represent a mix of Case 1 waters along the Portuguese coast, Celtic and Mediterranean seas and Case 2 waters of the English Channel and North Sea. The ability of the MEGS8.0 processor to switch between the bright pixel and clear water AC models, means that it can be applied to a diverse range of water types typical of NW Europe from the coastal Atlantic Ocean and North Sea. By comparison, the COASTCOLOUR AC showed a higher scatter than MEGS8.0 at coastal Atlantic, Mediterranean and English Channel stations (Figs. 3, 4). Using MEGS8.0 OC5Me in these NW European waters, the target error tolerance for Chl *a* is met (Table 2). These results have implications for the Sentinel-3 Ocean Colour Land Instrument (OLCI); the improved signal-to-noise ratio, long term radiometric stability and mitigation of sunglint and improvements in the AC for turbid and highly absorbing waters for OLCI, due to an increase in the number and position of spectral bands (Donlon et al., 2012), suggest that the accuracy of OC5 with OLCI could be improved further. Sentinel-3 OLCI will use the same or similar AC models that have been used for COASTCOLOUR and MEGS8.0 (Antoine 2010; Doerffer 2010; Moore and Lavender, 2010). Based on these results for MERIS, OC5 warrants further investigation with Sentinel-3 OLCI data.

Previous studies with SeaWiFS also found that accurate ocean colour estimates for the North Sea and English Channel can be achieved using the bright pixel AC in conjunction with band ratio Chl *a* algorithms (Blondeau-Patissier et al. 2004; Tilstone et al., 2013). This is facilitated by the use of the $R_{rs}(490): R_{rs}(560)$ ratio to derive Chl *a* (Tilstone et

al., 2013), since these bands are less affected by errors in the AC or due to high $a_{CDOM}(\lambda)$ absorption in the blue portion of the spectrum. Moore et al. (2009) assessed the uncertainty in Chl *a* in eight optical water types classified on the shape of $R_{rs}(\lambda)$, and found that only in clear waters was the 35% mission error met. In turbid high sediment water types, the relative error increased to >100%. By comparison with MODIS-Aqua, the largest errors were encountered for water types dominated by phytoplankton and a_{CDOM} (Goyens et al., 2013). Lee et al. (2010) used the theory of error propagation to derive the uncertainties in the inversion of inherent optical properties from $R_{rs}(\lambda)$ using the quasi analytical algorithm QAA (Werdell et al., 2013) with a simulated data set and found that the error in $a(440)$ was 13 to 37% over a range of $a_{CDOM}(440)$ from 0 to 2 m^{-1} . We therefore also addressed the question; what is the effect of $a_{CDOM}(442)$ and non-algal SPM on the Chl *a* algorithms? To assess this we plot *in situ* SPM and $a_{CDOM}(442)$ against the MEGS8.0 Chl *a*: *in situ* Chl *a* ratio for OC5Me, OC4Me, AP2 and GSM (Fig. 12). The horizontal dotted line represents where the algorithm Chl *a* equals *in situ* Chl *a* and data points above this line represent an over-estimate in algorithm Chl *a*, and points below the dotted line represent an under-estimate in Chl *a*. A significant correlation with SPM or $a_{CDOM}(442)$ would suggest that the under- or over-estimate in the Chl *a* algorithm could be partially accounted for by these variables. For OC5Me there was a significant positive correlation with non-algal SPM ($F_{1,69} = 6.91$, $P = 0.011$, $R^2 = 0.08$; Fig. 12A), suggesting that with MEGS8.0 the over-estimate in Chl *a* observed for this algorithm (Figs. 3A, 4A) was due to non-algal SPM. For GSM, there was a significant positive correlation between $a_{CDOM}(442)$ and algorithm: *in situ* Chl *a* ($GSM-F_{1,72} = 34.40$, $P < 0.0001$ $R^2 = 0.32$; Fig. 12H), suggesting that the over-estimate in Chl *a* (Fig. 4E) is due to $a_{CDOM}(442)$. For OC4Me, the over-estimate in Chl *a* at <10.0 mg m^{-3} (Fig. 4E) is due to non-algal SPM (Fig. 5A, D). The over-estimate in OC4Me at Chl *a* > 10.0 mg m^{-3} could be due to $a_{CDOM}(442)$, though there were not sufficient data points in Fig. 12D to verify this. There was no significant correlation between AP2 and SPM or $a_{CDOM}(442)$ suggesting that under- and over-estimate in AP2 Chl *a* is due to random error in the algorithm.

5. Conclusions

An accuracy assessment of MERIS Chl *a* in NW European waters, was conducted using MEGS8.0 and COASTCOLOUR processors with ocean colour algorithms that are widely available from the ESA NASA and NEODAAS. OC5Me Chl *a* was more accurate than OC3Me, OC4Me, GSM and AP2 Chl *a* using both COASTCOLOUR and MEGS8.0 processors, and MEGS8.0 OC5Me was slightly more accurate than COASTCOLOUR. Satellite images processed using MEGS8.0 $R_{rs}(\lambda)$ illustrated that OC4Me was 5 to 10 fold higher than OC5Me in coastal regions of the North Sea and English Channel, which was principally caused by errors in OC4Me that co-varied with SPM. The GSM was > 10 times higher than OC5Me in these regions, which was caused by variations in $a_{CDOM}(442)$. AP2 was the least accurate algorithm in these waters.

The error and uncertainty in both MEGS8.0 and COASTCOLOUR $R_{rs}(442)$ and $R_{rs}(560)$ over a Chl *a* range of 0.3 to 7 mg m^{-3} , were higher than the mission goal of 5% for the global ocean. The lower uncertainty in MEGS8.0 $R_{rs}(560)$ and the higher error in COASTCOLOUR $R_{rs}(560)$ led to slightly more accurate Chl *a* for the ocean colour algorithms tested with MEGS8.0. The performance of OC5 with MERIS data warrants further investigation with Sentinel-3 OLCI data for NW European and similar coastal and shelf waters.

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